# EBASCO

IEPA-93-32

April 7, 1993

Mr. Steven L. Gobelman, P.E. Remediation Engineering Sub-Unit Remedial Project Management Section Bureau of Land 2200 Churchill Road Springfield, Illinois 62794-9276

Subject:

Professional Service agreement No. BIE 9023

Sandoval Zinc - LPC #1210500002 Draft Feasibility Study Report

Dear Steve:

Please find enclosed for your review a copy of the Draft Feasibility Study Report for the Sandoval Zinc Site located in Sandoval, Illinois. The Executive Summary will be completed once approval of the report contents has been finalized. If you have any questions concerning the report please do not hesitate to call me.

Sincerely,

**EBASCO** Environmental

James Brinkman Senior Engineer

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#### 1.0 INTRODUCTION

The Illinois Environmental Protection Agency (IEPA) contracted with Ebasco Services Incorporated (EBASCO) to perform a Feasibility Study on the Sandoval Zinc Site located in Sandoval, Illinois, under the terms of State Multi-Site Professional Services Agreement Contract No. BIE-9023. The defined scope of work for this project included the following seven tasks:

Task 1 - Information Review

Task 2 - Information Summary

Task 3 - Data Needs and Assessment

Task 4 - Strategy Meeting

Task 5 - Feasibility Study

Task 6 - Topographic Map of Site

Task 7 - Project Task Control

After completing Task 4, EBASCO determined that there was insufficient information available to complete the feasibility study ("Information Summary and Data Assessment" Report; April 14, 1989) and recommended a supplemental field investigation. A Work Plan and Field Sampling Plan (dated June, 1989) were subsequently prepared for the supplemental field investigation.

This document is the feasibility study report based on all previously collected data and documents as well as the supplemental field investigation conducted by EBASCO. Sections 1 through 5 include a description of the investigation objectives, site background information, the study area, its physical characteristics (geology and hydrogeology), the nature and extent of impacted media, and conclusions drawn from the field investigation. These first five sections of the report cover the basic elements of a Remedial Investigation (RI) report. Sections 6 through 8 identify Applicable or Relevant and Appropriate Requirements (ARARs), screen relevant remedial technologies, develop appropriate remedial alternatives and compare the remedial alternatives. These last three sections of the report cover the basic elements of a Feasibility Study (FS). Section 9 lists the references used in preparing this report.

### 1.1 Field Investigation Objectives

The primary goals of the supplemental field investigation conducted by EBASCO in May and June, 1990, with a second phase sampling effort in April 1991, were as follows:

- O Develop specific information about the nature, extent, and level of contamination at the Sandoval Zinc site.
- O Determine the physical and chemical background characteristics of the soil and the chemical background characteristics of the groundwater.
- O Define the nature and extent of impacted soils, surface water, sediment, groundwater on-site.
- o Evaluate potential off-site contaminant pathways in soil, surface water, and groundwater that may affect public health and the environment.
- o Identify and evaluate potential alternatives for remediation.

### 1.2 Site Background

The Sandoval Zinc site (IEPA Site Inventory Number 1210500002) is an abandoned zinc smelter facility located southeast of the town of Sandoval in Marion County, Illinois (Figure 1-1). The site covers approximately 12 acres and is relatively flat, owing to the large quantity of artificial fill (metal-rich cinders from the smelting process) that was used to level the site's natural topography.

### 1.2.1 Site Description

The Sandoval Zinc site is comprised of two large abandoned buildings, an abandoned railroad tank car (also referred to as the above ground storage tank), old furnace building ruins, a "farm pond" to the east, and a marshy area to the west (Figure 1-2). The site is covered with grey cinder fill and little vegetation grows on the fill material. Surface water runs off into drainage ditches located east and west of the on-site buildings. Fill material also appears to be carried by surface water runoff past the property line and is accumulating in the field immediately south of the fence line. Since the site slopes several feet down towards the "farm pond", it likely also receives surface water runoff from the site.

### 1.2.2 Site History

The Sandoval Zinc smelter facility began operating as a primary zinc smelter some time between 1885 and 1890. Approximately twenty-five years later, in 1915, the operations were converted to secondary zinc smelting and the facility continued to operate in this manner until the facility was closed in the 1980s. On June 27, 1972, the plant was almost entirely

destroyed by fire. The buildings were rebuilt and the plant continued to operate until 1985. On December 19, 1986, the Sandoval Zinc Company was officially dissolved and the owners declared bankruptcy.

For the first 85 years of operation, the principal waste emissions from the plant were metalladen cinders and windblown ash. Large quantities of the cinders from the smelting process were used in constructing and surfacing secondary roads in the plant area and as fill material on the plant property. As a result, a layer of metal-rich cinders, ranging from 1 to 10 feet in thickness, now covers approximately 12 acres of the plant site.

The windblown ash from the smelter stack settled on the plant site and the surrounding farmland. Assuming the plant was fairly typical of secondary zinc smelters using retort processing, these air emissions were probably rich in heavy metals and ranged from 50 to 100 tons per year from the retort alone. Additional wind-borne emissions could have been generated from plant waste-handling procedures such as open storage of cinders and ash, and bulk storage of products (principally zinc oxide) in bins within plant buildings.

In compliance with air pollution control regulations, a scrubber was installed on the plant stack in 1970. Wastewater from the scrubber was dewatered in a seepage pit constructed on-site. This pit held the sludge from the process until it was removed for zinc reclamation. Another pit was used for the disposal of baghouse dust and floor sweepings. Based on the information available the exact locations of the pits are currently not known.

### 1.2.3 Previous Investigations

The Illinois State Water Survey (ISWS) and the Illinois State Geological Survey (ISGS) carried out geologic and groundwater studies at the site from 1975 to 1982. The final study report, entitled Retention of Zinc, Cadmium, Copper, and Lead by Geologic Materials<sup>1</sup>, was published in 1982. Forty-nine monitoring wells were installed on-site at thirty-six different locations during the study and provided the primary source of information for the site. The study described the geologic materials underlying the site as follows:

Peoria Loess
Roxana Silt
Berry Clay (Glasford Formation)
Hagarstown Member (Glasford Formation)
Glasford Till
Lierle Clay (Banner Formation)
Banner Till
Bond Formation Shale

During the study, soil samples were collected from a variety of locations across the site and from control borings located approximately three miles south-southwest of the site. The samples were analyzed for cadmium, copper, lead, and zinc. Background concentrations for the four heavy metals tested were 20 to 50 mg/kg for zinc, 0.04 to 1.5 mg/kg for cadmium, 10 to 30 mg/kg for copper, and 10 to 40 mg/kg for lead. Based on these background samples, there appeared to be no significant naturally occurring chemical variation with depth or between geologic unit boundaries. However, some zinc levels in isolated Pleistocene soils were higher than the established background levels.

According to the ISWS/ISGS report, the zinc processing waste covering the site varies widely in metals content but is generally rich in zinc, lead, copper, and aluminum. Cadmium was also detected in the soil samples collected. One sample of waste material at the site was 76 times the EP Toxicity Standard for lead. This large volume of material represents both a potential environmental hazard as well as a source of reclaimed metals. Typical weight percentages of the metals are 23% zinc, 3.8% aluminum, 2.5% lead, and 0.5% copper. These heavy metals have penetrated site overburden to depths of up to 28 feet.

Piezometric surfacewater maps constructed by the ISWS/ISGS suggest that the underlying till has an extremely low hydraulic conductivity. The Peoria Loess, Roxana Silt, and Berry Clay appear to allow the slow percolation and infiltration of contaminants downward; however, till units below the Sangamon Soil of the Berry Clay appear to be acting as an aquiclude to the further downward migration of contaminants.

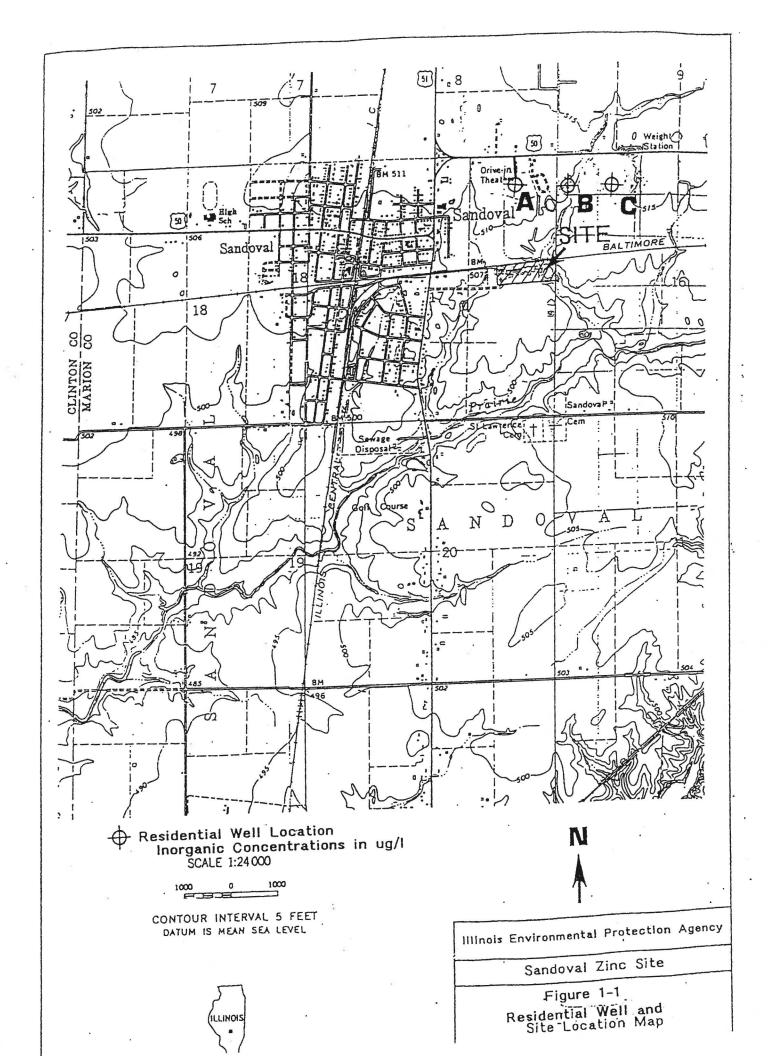
Laboratory analysis of groundwater samples from the monitoring wells sampled in 1975 and 1976 indicated that zinc contaminants had migrated from the wastes, through the soils, and into the groundwater of the Hagarstown Unit. Groundwater maps with contoured zinc concentrations were constructed for the ISWS/ISGS report and are presented in Figure 1-3. The maps show the extent of zinc impacted groundwater in August 1975 and September 1976 and indicate that the zinc plume is migrating from the source areas.

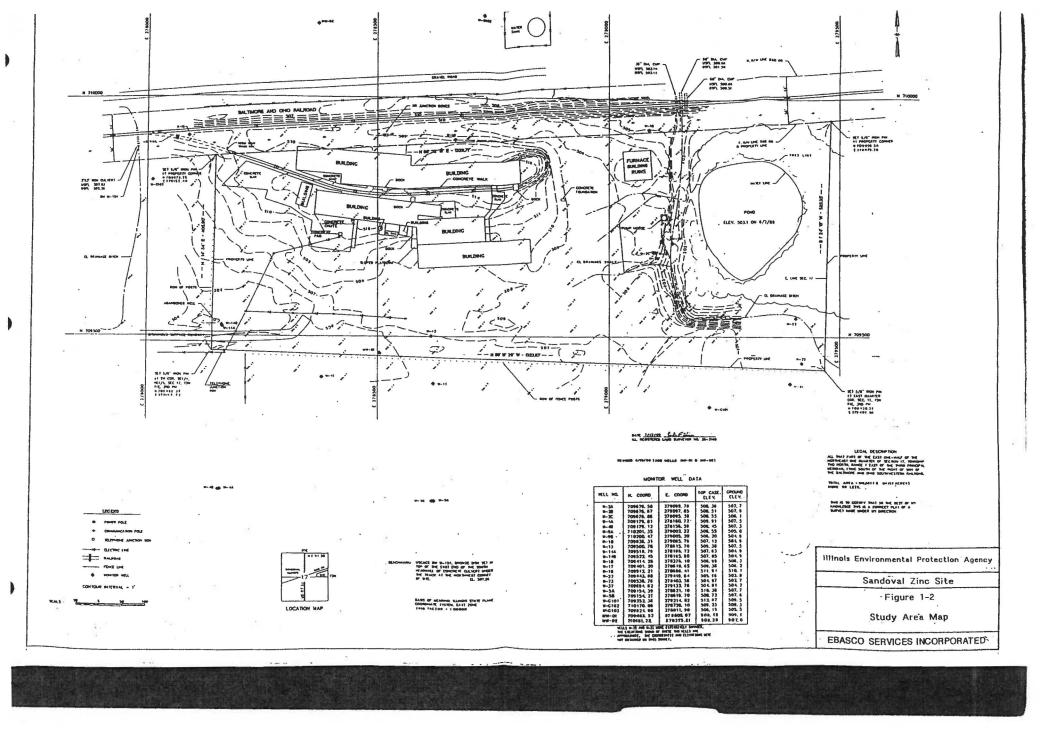
During 1986 and 1987, several sampling activities were conducted by IEPA and the Bureau of Mines at the Sandoval Zinc site. Soil, sediment, surface water, and groundwater samples were collected and analyzed for organic and inorganic parameters. The results of these sampling efforts are summarized in Tables 1-1 through 1-10. In March 1987, water and sediment samples were collected from the drainage ditches at the eastern and western edges of the site. Zinc and cadmium concentrations in the surface water samples exceeded the ambient surface water quality limits set forth in Section 304.124 of Subtitle C, IEPA's Water Pollution Regulations<sup>2</sup>. These limits are 1.0 mg/l and 0.15 mg/l for zinc and cadmium, respectively. High levels of these two heavy metals as well as other metals were detected in sediment samples collected downstream of the site. Zinc and cadmium concentrations

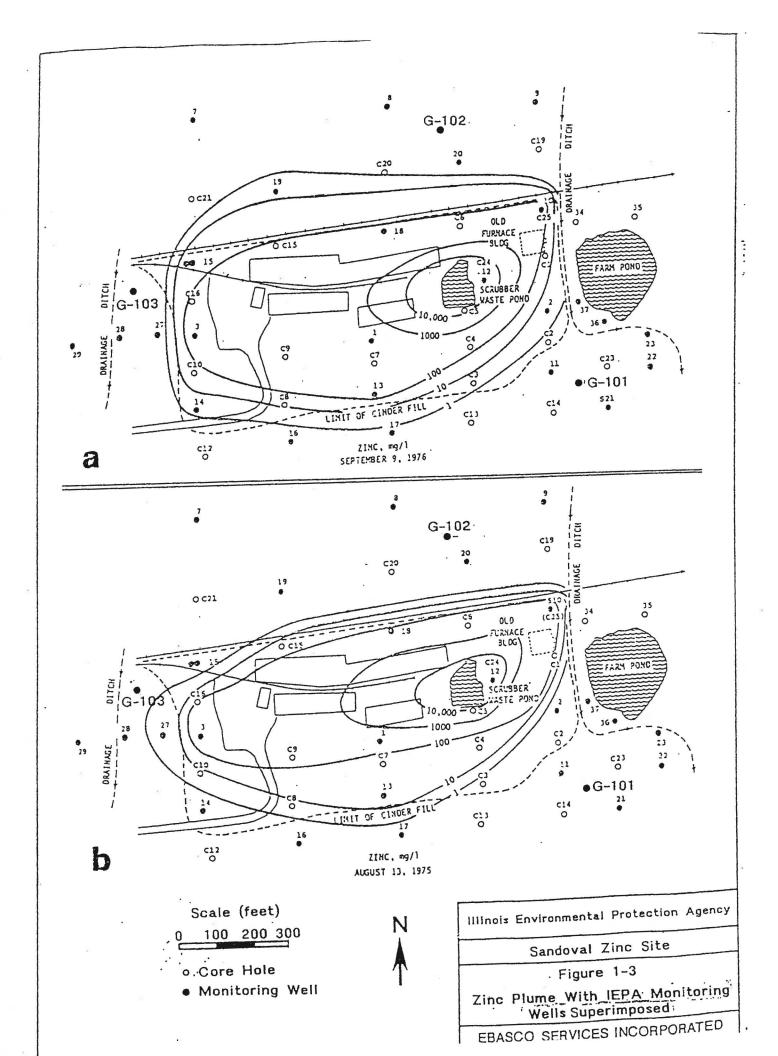
averaged greater than 17,000 mg/kg and 14 mg/kg, respectively in downstream sediments samples. The impacted waters and sediments in the drainage ditches could potentially reach Prairie Creek, approximately one-half mile from the site. However, the extent of migration of the waters and sediments is not fully known since sampling is relatively recent, and noncomprehensive in scope.

The ISWS/ISGS study identified that the primary mechanisms retaining the metals in the soils at the site were cation exchange and the precipitation of insoluble metal compounds due to changes in soil pH. Elevated levels of calcium and magnesium in groundwater samples during the IEPA studies of 1986 and 1987 suggested that cation exchange is continuing.

EBASCO performed a preliminary site visit on March 9, 1989. The purpose of the visit was to gather information necessary for preparing a Work Plan. During this visit, EBASCO made a preliminary assessments of sampling sites by matrix and location, determined the appropriate levels of personal protective equipment required, identified existing monitoring well locations and made an overall assessment of site conditions. On site and adjacent areas were visually inspected for contamination, including signs of surface water contamination, vegetation stress, physical hazards, and other environmental hazards. A complete description of the site visit and a photo log documenting those areas exhibiting signs of contamination are given in the Information Summary and Data Assessment Report<sup>3</sup>.







# Table 1-1 Summary of Analytical Results From Previous IEPA Study - Soil (ug/g) Sandoval Zinc Sandoval, Illinois

PARAMETER	X101	X102	X103	X201	X202	X203	X204	X205
Antimony	14	. 14	30	140	28	41	14	32
Arsenic	7	13	10	4	17	26	11	11
Beryllium	1	1	1	1	3	1	1	1
Cadmium	27.2	26.6	14.2	35.1	21.9	19.6	60.9	110
Chromium	39.1	18.1	23.5	1,360	9.8	9.2	22.9	22.2
Copper	1,240	418	880	34,200	320	1,780	1,560	2,810
Lead	7,560	1,590	5,650	25,800	40,000	10,000	69,600	22,400
Mercury	0.36	0.04	9.43	0.11	0.28	0.25	4.55	5.79
Nickel	570	14.4	230	12,100	52	220	610	800
Selenium	6	6	2	20	12	4	6	6
Silver	3.1	0.8	2.5	5.1	5.2	4.4	1.9	3.2
Titanium	3	3	2	2	6	2	3	3
Zinc	71,100	8,360	56,700	16%	40.3%	26,400	32.8%	19.5%

Note: Samples collected by IEPA on July 25, 1986.

# Table 1-2 Summary of Analytical Results Collected By IEPA Surface Water Samples (mg/l) Sandoval Zinc Sandoval, Illinois

PARAMETER	S101	S102	S103	S104	S105	S0106
Aluminium	0.33	0.51	0.54	0.84	1.9	2.2
Antimony	0.006	0.004	0.003	BDL <sup>1</sup>	0.003	BOL
Arsenic	BDL	BDL	BDL	BDL	0.022	BDL
Cadmium	0.510	0.465	0.59	0.021	0.59	0.13
Chromium	0.018	0.018	0.018	0.022	0.022	0.022
Copper	0.041	0.038	0.027	0.01	0.034	0.01
Iron	0.75	0.83	0.59	1.2	1.4	1.8
Lead	0.076	0.052	0.048	0.014	0.034	0.006
Mercury	BOL	BDL	BDL	BDL	BDL	0.0002
Nickel	0.15	0.11	0.09	BDL	0.07	BDL
Silver	0.008	0.007	0.01	0.007	0.01	0.008
Zinc	52	46	20	0.65	23	0.12

#### 1 - Below Detection Limit (BDL)

Note: Samples collected by Dennis Newman on March 30, 1987.

# Table 1-3 Summary of Analytical Results Collected By IEPA Sediment Samples (mg/kg) Sandoval Zinc Sandoval, Illinois

					<del>,</del>		
PARAMETER	X101	X102	X103	X104	X105	X106	X110
Aluminium	4,733	8,732	7,712	5,029	5,476	7,020	2,344
Antimony	1.6	0.26	BDL <sup>1</sup>	0.5	0.27	0.28	16
Arsenic	26	24	18	15	BDL	12	12
Cadmium	24	13	19	7.4	1.5	2.2	46
Chromium	34	41	22	25	11	18	20
Copper	1,065	960	252	688	38	71	1,250
Iron	20,710	17,990	7,854	8,621	5,476	9,200	5,370
Lead	710	1,660	15	1,026	189	140	1,760
Mercury	2.5	0.43	0.08	0.18	0.05	0.04	2.7
Nickel	716	515	117	287	5.4	28	114
Silver	2.2	2.5	1.4	1.6	1.0	1.1	8.0
Zinc	15,310	24,230	4,863	13,700	507	1,158	62,400

1 - Below Detection Limit (BDL)

Note: Samples collected by Dennis Newman on March 30, 1987.

# Table 1-4 Summary of Analytical Results Collected By IEPA Additional Soil Samples Sandoval Zinc Sandoval, Illinois

04	SAMPLE	X101	SAMPLE	X102	SAMPLE	X103
PARAMETER	Total (mg/kg)	E.P. Toxicity (mg/l)	Total (mg/kg)	E.P. Toxicity (mg/l)	Total (mg/kg)	E.P. Toxicity (mg/l)
Aluminium	6,030	0.64	5,860	1.03	1,930	0.68
Antimony	BDL <sup>1</sup>	BDL	BDL	BDL	69.8	BDL
Arsenic	28.1	BDL	38.5	BDL	25.2	BDL
Cadmium	23.3	0.5	60.3	1.0	7.6	0.1
Chromium	16.6	BDL	32.2	BDL	8.79	BDL
Copper	1,850	3.0	2,710	9.0	956	2.0
Iron	12,900	0.99	14,600	BDL	5,280	BDL
Lead	29,200	381	10,600	14.2	38,900	106
Mercury	8.8	BDL	1.3	BDL	1.2	BDL
Nickel	547	3.0	281	1.0	114	0.6
Silver	3.82	BDL	7.92	BDL	4.36	BDL
Zinc	226,000	2,400	281,000	1,800	493,000	2,200

1 - Below Detection Limit (BDL)

Note: Soil samples collected on April 9, 1987.

Table 1-5
Summary of E.P. Toxicity Results (mg/l) Collected By IEPA
Sandoval Zinc
Sandoval, Illinois

PARAMETER	X101	X201	X301
Aluminium	BDL <sup>1</sup>	BDL	BDL
Antimony	BDL	BDL	BDL
Arsenic	BDL	BDL	BDL
Cadmium	0.007	0.011	0.02
Chromium	0.01	0.02	0.01
Copper	BDL	BDL	BDL
Iron	1.1	0.13	0.1
Lead	0.009	0.018	BDL
Mercury	BDL	BDL	BDL
Nickel	BDL	BDL	BDL
Silver	BDL	BDL	BDL
Zinc	0.23	0.005	0.097

<sup>1 -</sup> Below Detection Limit (BDL)

Note: Samples collected by Kevin Rodgers on April 27, 1987.

Table 1-6
Summary of Analytical Results Collected By Bureau of Mines
Sandoval Zinc
Sandoval, Illinois

Analyte	Concentration (Weight Percent)
Aluminum	14.9%
Carbon	9.0
Iron	2.8
Lead	2.5
Silicon	14.9
Zinc	23.0

Sample was a composite soil sample collected by R.L. Johnson on February 20, 1987.

Note:

# Table 1-7 Summary of Organic Analytical Results Collected by Environdyne - Groundwater (ug/l) Sandoval Zinc Sandoval, Illinois

PARAMETER	G101	G102	G103
Diethylphthalate	2	7	-
Di-N-Butylphthalate	2	2	2
Hexanedioic Acid, Dioctylester	62	98	•
Carbon Disulfide	-	2	-
Benzene	E 6.4	9 3	1
Di-N-Octylphthalate	•	3	
2 (3H) Furanone, Dihydro	•	9	-
Hexanoic Acid, 6-Amino	<u>-</u>	12	-
Unknown	II <u>.</u>	4	-
Unknown	-	-	10
Unknown	-	-	7
Unknown		•	44

Note: Groundwater samples collected May 14, 1987

Table 1-8
Summary of Inorganic Analytical Results
Collected by Environdyne - Groundwater (ug/l)
Sandoval Zinc
Sandoval, Illinois

	<del>7</del>	<del></del>	<del></del>
PARAMETER	G101	G102	G103
Aluminium	BDL <sup>1</sup>	BDL	82
Antimony	BDL	BDL	BDL
Arsenic	BDL	BDL	BDL
Cadmium	BDL	BDL	BDL
Chromium	BDL	BDL	BDL
Copper	BDL	BDL	BDL
Iron	140	BDL	61
Lead	2	BDL	BDL
Mercury	BDL	BDL	BDL
Nickel	BDL	BDL	BDL
Silver	BDL	BDL	BDL
Zinc	36	24	110
Sulfate	276,000	89,600	273,300
Sulfide	BDL	BDL	1,600

1 - Below Detection Limit (BDL)

Note: Groundwater samples collected May 14, 1987

Table 1-9
Summary of Organic Analytical Results
Collected by Environdyne - Soil (ug/kg)
Sandoval Zinc
Sandoval, Illinois

	<del> </del>			
PARAMETER	S101	S102	S201	S202
Acetone	63	160	76	
2-Butanone	•	21	7	
2 (3H) - Furanone, Dihydro	•	1,856	1,608	1,831
Toluene		15	•	
Chloroform	-	• 12	-	6
Naphthalene	-	-	-	190
Dibenzofuran	•	-		190
Phenanthrene	•	(-	-	1,00
Anthracene	•			220
Fluroanthene	•	-	-	1,100
Pyrene	•	•	•	1,400
2-Methylnaphthalene		•	-	900
Benzo (a) Anthracene	-	-	-	570
Chrysene	-	•	•	1,300
Benzo (b) Fluoranthene	-	-	-	710
Jnknown Organics <sup>1</sup>		40,000	45,000	24,000

Note: Samples collected May 14, 1987

# Table 1-10 Summary of Inorganic Analytical Results Collected by Environdyne - Soil (ug/kg) Sandoval Zinc Sandoval, Illinois

PARAMETER	S101	\$102	S201	S202
Aluminum	10,500,000	15,300,000	15,200,000	10,300,000
Antimony	43,600	BDL	BDL	BDL
Arsenic	27,500	14,300	5,670	4,900
Cadmium	49,000	44,600	23,200	11,200
Chromium	18,600	14,400	16,600	41,400
Copper	446,000	129,000	67,200	1,370,000
Iron	32,100,000	15,300,000	18,500,000	41,600,000
Lead	1,226,000	272,000	139,000	4,662,000
Mercury	BDL	BDL	BDL	670
Nickel	199,000	40,000	20,600	334,000
Silver	BDL	BDL	BDL	1,500
Zinc	10,300,000	6,030,000	3,770,000	44,700,000
Cyanide	350	BDL	BDL	BDL
Sulfate	?	2,600	7	110,200
Sulfide	BDL	7,600	34,200	BDL

Note: Samples collected May 14, 1987

#### 2.0 STUDY AREA INVESTIGATION

This section presents the scope of the supplemental field investigations and describes how each component of the investigation was conducted.

### 2.1 Scope of Supplemental Field Investigation

The supplemental field investigation effort at the Sandoval Zinc site focused on collecting the data needed to sufficiently characterize the site in order to evaluate and select remedial actions that would adequately protect human health and the environment. Prior to beginning the field activities, a Work Plan, including a site-specific Field Sampling Plan (FSP), and a Health and Safety Plan (HASP), were developed. A detailed Quality Assurance Project Plan (QAPP) was not developed because all analytical work was performed by an IEPA approved laboratory (ARDL Laboratories) participating in the Contract Laboratory Program (CLP). Additional documents related to the supplemental field investigation and the surveying subcontract were also prepared. The work plan explains the purpose for each component of the investigation including number of samples, locations and analytes.

The EBASCO field activities at the Sandoval Zinc site were conducted from May to June 1990. The investigation included air monitoring, surface soil and sediment sampling, borehole drilling and monitoring well installation, permeability testing at selected monitoring wells, residential well and groundwater sampling, surface water sampling, sampling of waste product and ash from the interiors of the buildings, and sampling the contents of the abandoned above ground storage tank.

### 2.2 Topographic Survey

A site survey encompassing approximately twelve acres was conducted in June 1989. The final survey map produced includes the natural features and permanent structures located on-site. Also included on the map are ground surface elevations, property boundaries, the locations of the wells and cores of the ISWS/ISGS investigation (where possible), and the location and extent of the "farm pond". The surveying activities were performed by Hanson Engineers, Incorporated, of Springfield, Illinois. Coordinates on-site were established from the Illinois State Plane Coordinate System, East Zone. Four points were set on or near the site as baseline points, two located on the B&O railroad tracks at the northern boundary of the site, and the remaining two points on-site. Elevations were established from a benchmark at the northwest corner of the site and are based on the National Geodetic Vertical Datum (NGVD) of 1929. Supplemental elevations were established as reference

points on the Sandoval Water Tank to the north of the site, the four control points, and the tops of monitoring well protective casings, where useable.

Hanson Engineers submitted a report entitled <u>Final Report of Survey Activities</u>, <u>Sandoval Zinc Site</u>, <u>Sandoval</u>, <u>Illinois</u><sup>4</sup> to EBASCO in July 1989. A copy of the final topographic survey map is provided in Appendix A.

### 2.3 Aboveground Tank Investigation

The tank investigation was conducted to identify the contents of the abandoned railroad tank car. The tank car is located on the railroad spur at the south side of the westernmost building on-site (Figure 2-1). One composite sample and a duplicate sample were collected from the tank using the sampling procedures outlined in the FSP. EBASCO personnel performed the sampling using Level C protective equipment.

The tank contents were visually inspected for stratification prior to sampling. Clear tubing was lowered into the tank as far as possible and withdrawn. The liquid in the clear tubing was determined to be oil, and was not stratified. The depth of the oil in the tank was approximately three feet eleven inches. The samples collected were analyzed for the volatile organic compounds benzene, toluene, ethyl benzene, and total xylenes (BTEX), the Target Analyte List (TAL) inorganics, pesticides and PCBs, and for heating value.

During the winter of 1991, a valve in the tank piping failed and released all of the residual liquid in the tank. As a result the IEPA conducted an emergency response action and removed approximately 500 cubic yards of petroleum impacted soil. The soil is presently stored inside one of the buildings on-site and the tank is now empty.

### 2.4 Waste Product/Ash Investigation

The waste product/ash investigation was designed to characterize the ash and waste product that is located in the buildings on-site. Some of the material was in labelled bags (zinc oxide, rock salt), but the majority of the waste product and ash in the buildings had been left in uncovered piles or was scattered across the floors. Composite samples were taken where distinct piles of waste product existed, otherwise samples were collected from scrapings off the floor.

Six samples and one duplicate sample of the waste product and ash were collected from various locations inside the buildings (Figure 2-2). WPA01S and the duplicate WPA01D were collected from an unlabelled bag containing waste product. WPA02S was a composite sample from eight discrete locations within a pile of waste product. Waste product from

three sides of an abandoned blower inside the building formed the composite sample WPA03S, and WPA04S was collected from a patch of discolored soil in the doorway to the oil tank. Sample WPA05S was a composite sample collected from various locations around the structures in the center of the building. The final composite sample of waste product/ash (WPA06S) was collected from a location near the door to the building. All samples were analyzed for full TAL inorganics and EP Toxicity.

### 2.5 Surface Soil and Sediment Investigation

The soil investigation was designed to establish the extent of shallow (less than 1 foot) surface contamination at the Sandoval Zinc site. Surface soil samples were collected from the locations shown in Figure 2-3 and were analyzed for full TAL inorganics. Selected samples were also analyzed for pesticides/PCBs. All surface soil and sediment samples were collected using a garden trowel and were typically collected from the top six inches of the surface soils.

Twenty-three surface soil samples and three duplicate samples were collected to give the most coverage to the surface soil characterization. Some of the samples were taken off-site, from the northern side of the railroad tracks, and outside the southwestern site boundary, and the remaining samples were collected from locations at random across the site and adjacent to the site buildings.

Four sediment samples and one duplicate sample were collected from the perimeter of the farm pond to characterize the surface sediments in this area. SS01S was collected in the drainage ditch at the eastern portion of the site that drains into the farm pond (Figure 2-3). SS02S was a composite sediment sample collected from the western half of the farm pond, and SS03S was a composite sample from four locations on the eastern part of the pond. The last sediment sample, SS04S, was collected from the floodplain area southeast of the farm pond. All sediment samples were analyzed for TAL inorganics and pesticides/ PCBs.

### 2.6 Surface Water Investigation

The surface water investigation was conducted to characterize the waters of the "farm pond" and in the drainage ditch on the eastern side of the site. The four surface water samples and a duplicate sample were collected from the locations shown in Figure 2-4. Samples were transferred directly to the sample bottles and then labelled for the appropriate analyses.

Surface water sample SW01S and the duplicate SW01D were collected from the water in the eastern drainage ditch. Surface water sample SW02S was collected from the western

half of the farm pond, and SW03S from the eastern half of the pond. Sample SW04S was collected from standing water in a depression east of the farm pond. The samples were all analyzed for full TAL inorganics and the volatile organic compounds benzene, toluene, ethyl benzene, and xylene (BTEX).

### 2.7 Groundwater Investigation

The groundwater investigation was designed to determine the nature and extent of impacted on-site shallow groundwater and possible off-site impacted groundwater. The groundwater investigation consisted of installing two new monitoring wells, locating and assessing the conditions of existing monitoring wells, field permeability testing on selected monitoring wells, and groundwater sampling of five on-site wells and one off-site residential well.

Previous investigations at the Sandoval Zinc site had included the installation of numerous monitoring wells. The Illinois State Water Survey (ISWS) and the Illinois State Geological Survey (ISGS) installed 49 monitoring wells at 36 locations during their study; the Illinois EPA installed three monitoring wells (G101, G102, G103) at the site in 1987. EBASCO performed a site visit in June 1989 and attempted to determine the locations and conditions of these existing monitoring wells. Of the 49 wells installed for the ISWS/ISGS study, only 21 were located. Of these 21, only 13 were usable, and only for obtaining water level measurements. The three monitoring wells installed by IEPA were in good condition and were usable for water level measurements, groundwater sampling, and permeability testing.

The two shallow monitoring wells (MW01 and MW02) were installed during EBASCO's field activities of May and June, 1990. The locations of the two new monitoring wells and the three IEPA monitoring wells are shown in Figure 2-5. The newly installed monitoring wells were completed to depths of approximately 20 feet below ground surface, at the bottom of the Hagarstown aquifer unit. The three IEPA wells were all completed at a depth of approximately 17 feet below ground surface. Information on these five monitoring wells, including installation dates, total depths, screened intervals, and completion zones is given in Table 2-1.

The two monitoring wells were installed using 3 and 3/4-inch inside diameter (I.D.) hollow stem augers with 3-inch I.D. continuous split-spoon samplers. The wells were constructed with 2-inch I.D. stainless steel well casings and risers. The 5-foot long stainless steel well screens had slot sizes of 0.010 inches. A minimum of one foot of sand was put in each borehole before the well casing and screen were lowered down. The sand pack extended approximately two feet above the top of the screen, and a two-foot seal of 1/2-inch diameter bentonite pellets was installed above the sand pack. Cement-bentonite grout was then

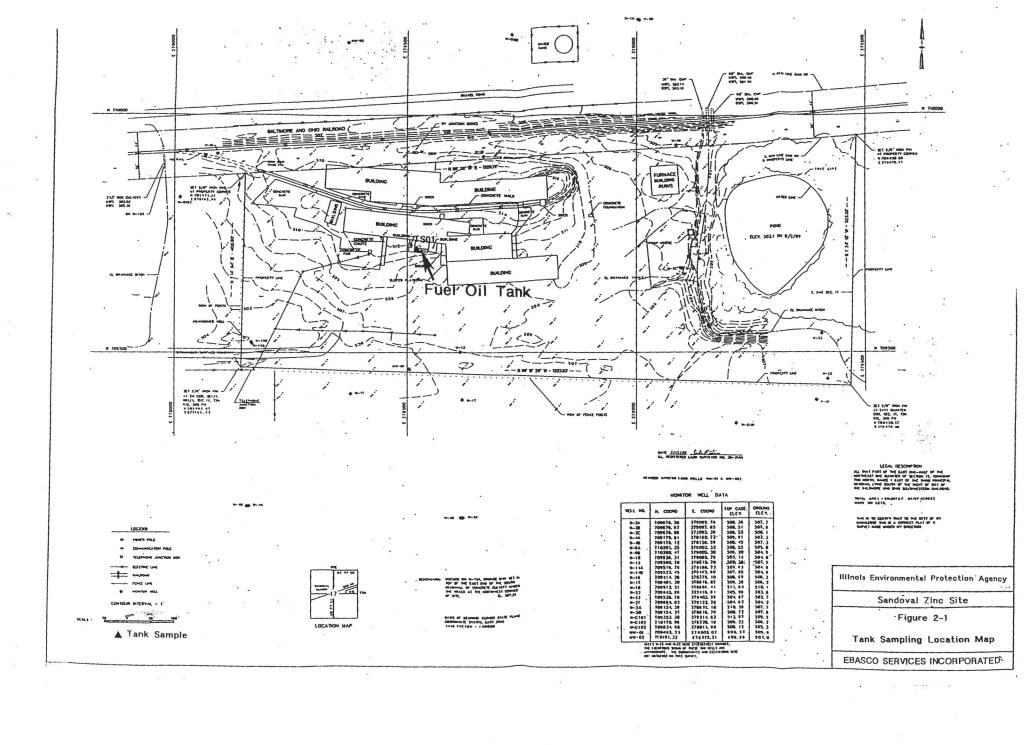
added to the surface, and the protective casing installed. Figure 2-6 presents a typical monitoring well construction diagram.

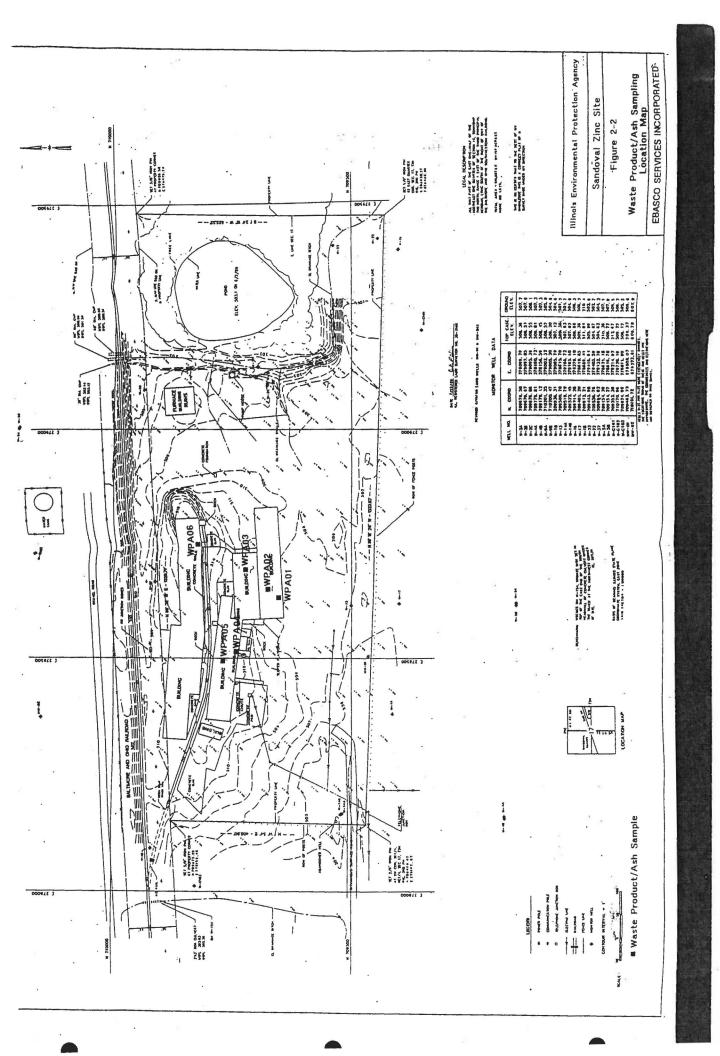
The monitoring wells were developed by bailing after a 24 hour stabilization period. Development continued until the parameters (temperature, pH, and conductivity) had stabilized and/or a sufficient well volume was purged so that the water was clear. The water purged from the wells was routed to the nearest surface drainage ditch for disposal.

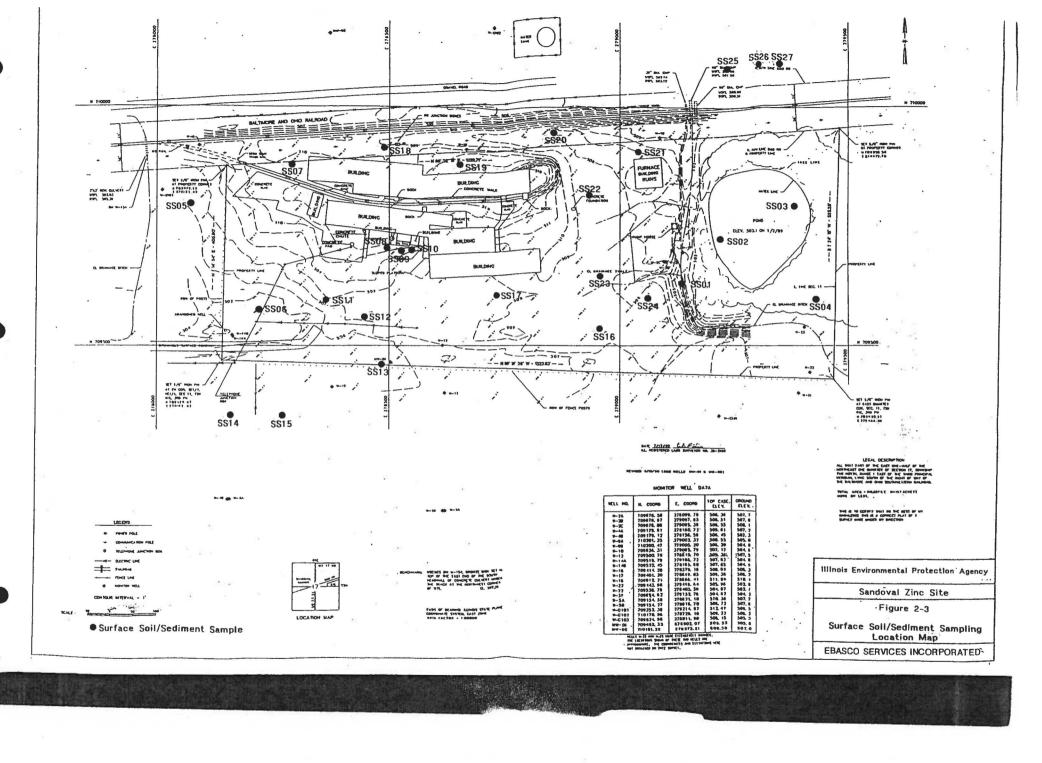
Slug tests consisting of falling and rising heads were conducted on the IEPA monitoring wells (G101, G102, and G103) in May 1990 and on the two newly installed monitoring wells in June 1990. The slug tests were performed by first measuring the static water level with an electronic tape. Then a 4-foot long, 1 and 1/4-inch outside diameter (O.D.) stainless steel slug was instantaneously lowered into the water until it was fully submerged. The water level drop was measured at timed intervals and recorded using a pressure transducer and data logger. The test continued until the water level in the well stabilized. The rising head test immediately followed, when the slug was removed from the well, and the water level rise was recorded. The data from all the slug tests are provided in Appendix B.

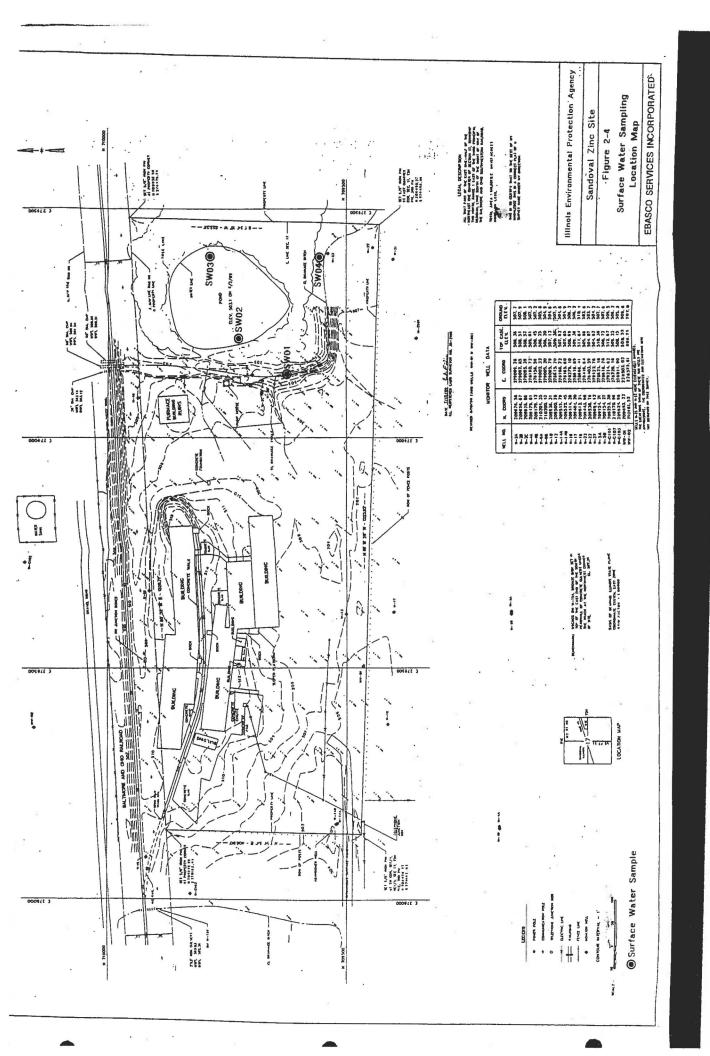
Groundwater samples were collected from the IEPA monitoring wells in May 1990 and from the two new monitoring wells (MW01 and MW02) in June 1990. Sampling at the new monitoring wells was done two weeks after well development was complete. Before the samples were collected, the depth of the water in each well was measured. The depth of the bottom of the well was noted, and the volume of standing water in the well calculated. Three well volumes of water were removed using a stainless steel bailer. The pH, conductivity, and temperature of the groundwater were recorded prior to sampling. The groundwater samples were collected using a stainless steel bailer and poured directly into the appropriate containers. Samples were then sent to the IEPA contract laboratory for full TAL inorganics, pesticides/PCBs, and BTEX volatile organics analyses.

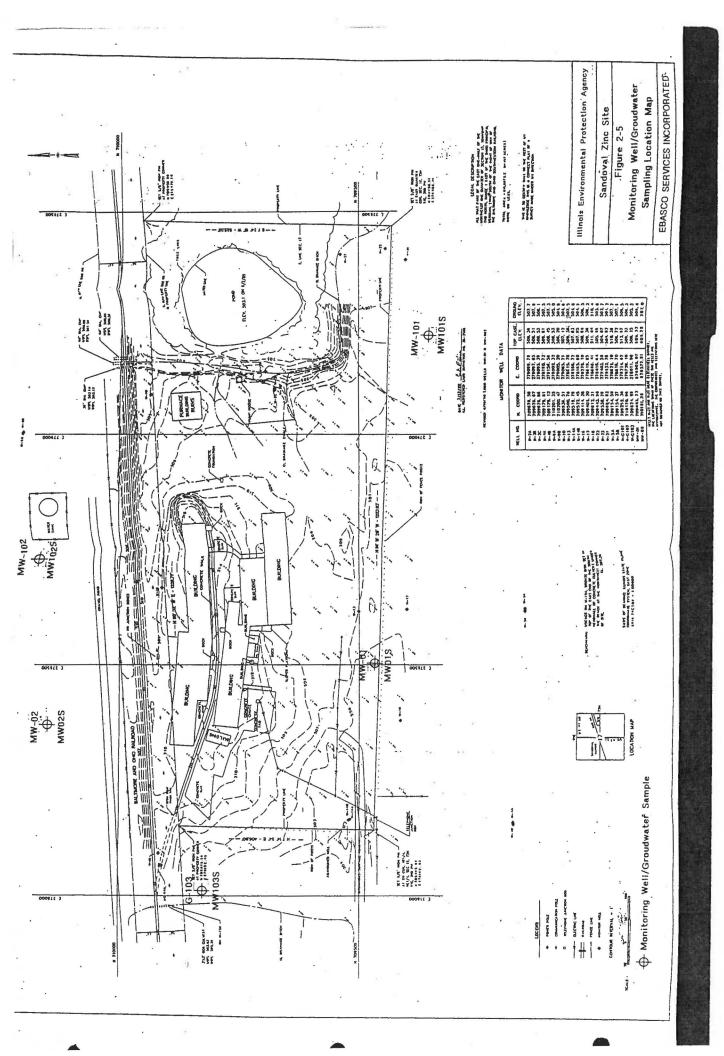
In the approved Work Plan, groundwater sampling was proposed for two of the residential wells located within a one mile radius of the site. During the field investigation of June 1990 only one residential well was located and sampled (Figure 1-1). The groundwater sample and duplicate were analyzed for full TAL inorganics.











## MONITORING WELL INSTALLATION

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EBASCO SERVICES INCORPORATED

Table 2-1 Monitoring Well Data Sandoval Zinc Sandoval, Illinois

WELL <sup>1</sup>	MW101	MW102	MW103	MW01	MW02
Elevation of Ground Surface (ft)	509.5	506.3	505.3	505.8	507
Elevation of Top of Casing (ft)	512.67	509.23	508.15	509.33	509.33
Depth to Top of Formation (ft)	11.2	12.2	12.3	12.5	14.4
Depth to Water (6/28/90) (ft)	7.34	4.74	4.74	4.44	4.42
Depth to Top of Screen (ft)	11.7	12	12	10.76	13.35
Elevation of Piezometric Surface (ft) (6/28/90)	505.33	504.49	504.42	504.89	504.96
Elevation of Top of Screen (ft)	497.8	494.3	493.3	495.04	493.65
Elevation of Top of Formation (ft) (Hagerstown)	498.3	494.1	493.3	495.04	492.6
Well Completion Data	4/06/87	4/08/87	4/06/87	6/14/90	6/15/90

<sup>1 - 5</sup> foot screen

### 3.0 PHYSICAL CHARACTERISTICS OF THE STUDY AREA

This section presents the physical characteristics of the study area based on available literature, data from previous investigations, and the additional field activities conducted by EBASCO for the IEPA. Special emphasis was given to those features needed to characterize the site for use in evaluating and selecting remedial alternatives.

### 3.1 Demography and Land Use

The Sandoval Zinc site is located southeast of the City of Sandoval in Marion County, Illinois (Figure 3-1). The 1980 census records list the total population of Marion County at 41,561, yielding an estimated 75 people per square mile in the county. The 1980 census also indicates that the population of Sandoval, Illinois was 1,535 people; this extrapolates to approximately 240 people living within a 1-mile radius of the Sandoval Zinc site.

The land area in the immediate vicinity of the Sandoval Zinc site is used for a variety of purposes. The land immediately south of the site is farmland, and the land north of the site is undeveloped grassland. West of the site, along Route 51, are several small businesses, and adjacent to Sandoval Zinc on the west is a junkyard and scrap metal yard. During the field investigations of May and June 1990, several piles of trash and tires in the junkyard appeared to be smoldering. There are marshy areas on the eastern and western edges of the site, and building refuse and scrap are littered across the site.

#### 3.2 Topography

Marion County is located in south-central Illinois in the physiographic region known as the Springfield Plain (ISGS, 1975). The land surface has been modified by glacial activity into the relatively flat to gently rolling plains characteristic of glacial drift regions. Surface elevations in the county range from approximately 475 to 520 feet above mean sea level (MSL).

The topography in the vicinity of the site is relatively flat and lies at approximately 500 to 505 feet MSL. An artificial mound of cinder and other fill material has raised the elevation of the central part of the site to approximately 510 feet MSL. The surface elevation of the farm pond at the eastern site boundary was surveyed in 1989 at 503.1 feet MSL. The site surface slopes gently to the lower elevations on all sides, except to the east, where a rapid drop of about 5 to 8 feet occurs, down to the farm pond. A topographic map with ground surface elevations at the one-foot contour interval is shown in Figure 3-2.

#### 3.3 Surface Water

The Sandoval Zinc site is located within the Prairie Creek drainage basin. Prairie Creek, which is the nearest surface water body in the vicinity of the site (Figure 3-1), flows to the south west about one half mile south of the site. Approximately six miles south-southeast of the Sandoval Zinc site is the Centralia Reservoir and Crooked Creek.

Surface water runoff at the site is controlled by site topography and the existing drainage ditches to the east and west. Since the central part of the site is the highest topographically due to the artificial fill, surface water runoff is in all directions away from the buildings. Surface water drains into both ditches, but primarily into the eastern ditch near the farm pond. Runoff from the site likely carrys material south, away from the site and into the neighboring field.

### 3.4 Geology

## 3.4.1 Regional Geology

The Sandoval Zinc site is located in the south central portion of the Illinois Basin, a large Paleozoic spoon-shaped sedimentary basin. Surficial deposits overlying the bedrock strata of the basin are unconsolidated glacial tills, outwash, and drift. The thickness and composition of these glacial deposits vary across the state, typically thinning to the south (Willman et al., 1975)<sup>5</sup>. Figure 3-3 is a generalized stratigraphic column of Pennsylvanian and younger sediments of south central Illinois.

The glacial deposits of south central Illinois are composed primarily of till, poorly sorted clay, silt, sand, and pebbles laid down during the four major Pleistocene advances of the glaciers (the Nebraskan, Kansan, Illinoisan, and Wisconsinan glacial advances). The periods of time between the glacial advances were known as the interglacials, and were times of soil formation (the Aftonian, Yarmouthan, and Sangamonian interglacials).

The Nebraskan and Kansan glacial advances represent the first two episodes of Pleistocene glaciation in Illinois. The Nebraskan glacial advance effected a small portion of western Illinois and was either never deposited in south central Illinois or subsequently eroded. In areas where Nebraskan glacial deposits occur it is common to find the Afton Soil formed on top of the deposits. The Kansan glacial advance effected nearly two-thirds of Illinois. Sediments deposited during the Kansan glacial advance in south central Illinois belong to the Banner Formation Till and the Lierle Clay Member overlies the till of the Banner Formation. The Yarmouth Soil was developed directly on top of the Kansan glacial deposits

during Yarmouthian time. The Lierle Clay units is part of the Yarmouth Soil but is an accretionary deposit made largely throughout Yarmouthian time (Willman et al., 1975).

The Illinoisan stage was marked by three major glacial advances into which covered most of the state. The Glasford Formation Till was deposited during the first and second glacial events of the Illinoisan stage. The Vandalia Till Member of the Glasford Formation was deposited during the second glacial event of the Illinoisan stage since the ice sheet stopped well north of south central Illinois during the final phase of glaciation. The Hagarstown Member of the Glasford Formation was then deposited. The Berry Clay Member of the Glasford Formation has been identified as a Sangamon accretion gley (Willman and Frye, 1970)<sup>6</sup>. Sagamon Soil developed directly on top of the Illinoisan deposits.

There were two glacial advances during the Wisconsinan stage. Wisconsinan glacial deposits were limited to northern Illinois, with large quantities of loess deposited over much of the rest of the state. Roxana Silt, a loess was deposited during the early and middle Wisconsinan during the first of the two glacial advances. The Farmdale Soil was a result of an interval of soil formation between the two Wisconsinan advances. Peoria Loess was then deposited as the result of deflation of alluvial deposits from outwash streams of late Wisconsinan glaciers.

The regional framework of bedrock strata in Illinois is controlled by the Illinois Basin. Strata underlying the study site range from Pre-Cambrian granites (oldest) to Pennsylvania sedimentary layers (youngest). The strata generally strike northeast and dip and thicken to the southeast, towards the center of the basin. The Pre-Cambrian basement rocks in Illinois are granites and granodiorites. They lie at depths greater than 8,000 feet below the ground surface in Marion County, and deep well investigations have shown up to several hundred feet of variation in the surface layer of these Pre-Cambrian rocks.

The preglacial bedrock surface in Marion county, Illinois belongs to the Pennsylvanian Bond Formation. These Pennsylvanian rocks consist predominantly of green calcareous clays and shales interbedded with thin sandstone, limestone, and coal layers. The Bond Formation varies from less than 150 feet thick in eastern Illinois to over 300 feet in southeastern Illinois and is approximately 250 feet thick in much of Marion County.

## 3.4.2 Site Geology

The subsurface geology at the Sandoval Zinc site was interpreted from EBASCO boring logs and previously existing boring logs of the IEPA. Two generalized cross sections were constructed from these logs. The locations of the cross-sections are shown in Figure 3-4. One cross-section was north-south (Figure 3-5), and the other was east-west (Figure 3-6).

The depths and thicknesses of the subsurface strata indicated were generalized from and interpreted between the borings. Information on actual subsurface conditions exists only at the locations of the well borings. Monitoring well boring logs and well construction diagrams can be found in Appendix B.

The generalized stratigraphy at the site, beneath the artificial cinder fill, consists of glacial deposits of varying thickness overlying the Pennsylvanian Bond Shale. From the EBASCO and IEPA boring logs, the glacial deposits, to depths of approximately 20 feet below ground surface, consist of the Peoria Loess and the Roxana Silt of the Wisconsinan Glacial Stage; the Berry Clay of the Sangamonian Stage; and the Illinoisan Stage Hagarstown Member and the Glasford Till. The Peoria Loess is a brownish-grey clayey silt with small amounts of sand (ISWS/ISGS, 1982) that was formed by wind deposits of fine particulate matter. The loess ranges in thickness from 6 to 12 feet across the Sandoval Zinc site. The Roxana Silt is described as a dark brown clayey silt with a fair percentage (20-34%) of sand. The Roxana Silt is thin underneath the site, thickness range from 1 to 2 feet. The Berry Clay is distinguished from the overlying silt by its dark-grey color and texturally it is a sandy, silty clay with some gravel (ISWS/ISGS, 1982). The Hagarstown Member of the Illinoisan Stage is a thin (1 to 2 foot) silty sand, that is variable in both thickness and composition; at times it is difficult to distinguish from the underlying till. The Hagarstown is the only unit which is water-bearing in the vicinity of the site. The Glasford Till consists of grey to dark grey sandy and silty till. Small lenses of sand, silt, and clay can be found within the till, which has thicknesses of approximately 20 to 40 feet.

Previous investigations by the ISWS/ISGS determined the glacial deposits below the Glasford Till to be the Lierle Clay and the Banner Formation Till. Underlying the Banner Formation Till, at depths of 55 to 75 feet below ground surface is the Pennsylvanian Bond Formation, a micaceous green shale. The EBASCO and IEPA borings were shallower than the borings of the ISWS/ISGS study, and were also located at the edges of the site, where the artificial fill material was not encountered.

#### 3.5 Groundwater

## 3.5.1 Regional Groundwater

Much of the regional groundwater in Marion County, especially in the western portion of the county, is retrieved from the unconsolidated glacial deposits that cover the Pennsylvanian bedrock. In limited areas, Pennsylvanian sandstones are a source of groundwater, especially in the southeastern portion of the county. Where the sandstones occur, groundwater can be recovered from the top 150 to 200 feet of the units (ISGS, 1957).

A buried valley is present in the west central part of Marion County. The pre-glacial valley has thick deposits of unconsolidated materials, especially sand and gravels. Buried valleys in the county to the west of Marion County is also a potential source of private and municipal water supplies.

#### 3.5.2 Local Groundwater

## 3.5.2.1 Groundwater Availability

Most of the local water supply for the City of Sandoval and the surrounding farms is obtained from large-diameter wells completed in the unconsolidated deposits of the Hagarstown Member. These wells, which were either dug or bored, usually tapped lenses on thin layers of water-bearing silt sand or gravel only a few inches thick (ISWS, 1980). The wells range in depth from 30 to 60 feet and water levels may vary up to 10 feet due to seasonal precipitation and recharge changes. These wells typically produce only a few hundred gallons of water a day and offer no potential for providing a municipal supply. Test holes drilled into the underlying shale bedrock have yielded only a few thin beds of water-yielding sandstone and creviced limestone. Below depths of 100 to 150 feet, the water is likely to be too brackish for domestic use.

#### 3.5.2.2 Groundwater Elevation

Water level data from the EBASCO and IEPA monitoring wells completed in the Hagarstown Member are presented in Table 3-1. Figure 3-7 presents the groundwater elevation in Hagarstown Member based on the average water levels measured from May 1990 and June 1990. It appears the groundwater in the Hagarstown Member is under confined or semiconfined conditions. The general direction of groundwater flow in the Hagarstown is somewhat difficult to determine. In 1975 and 1976, the ISWS/ISGS study discovered that the groundwater formed a mound under the Sandoval Zinc site, a mound centered on the site buildings. It was thought at that time that the mound existed due to liquid disposal practices at the site during operation and the high permeability of the fill material.

Water level elevations taken during the field investigation in May and June 1990 were taken only from five wells. Many of the wells of the ISWS/ISGS study were either not located or found to be unusable. Water level data collected during this investigation is insufficient to determine the presence or absence of the groundwater mound reported in 1975 and 1976.

## 3.5.2.3 Hydraulic Conductivity

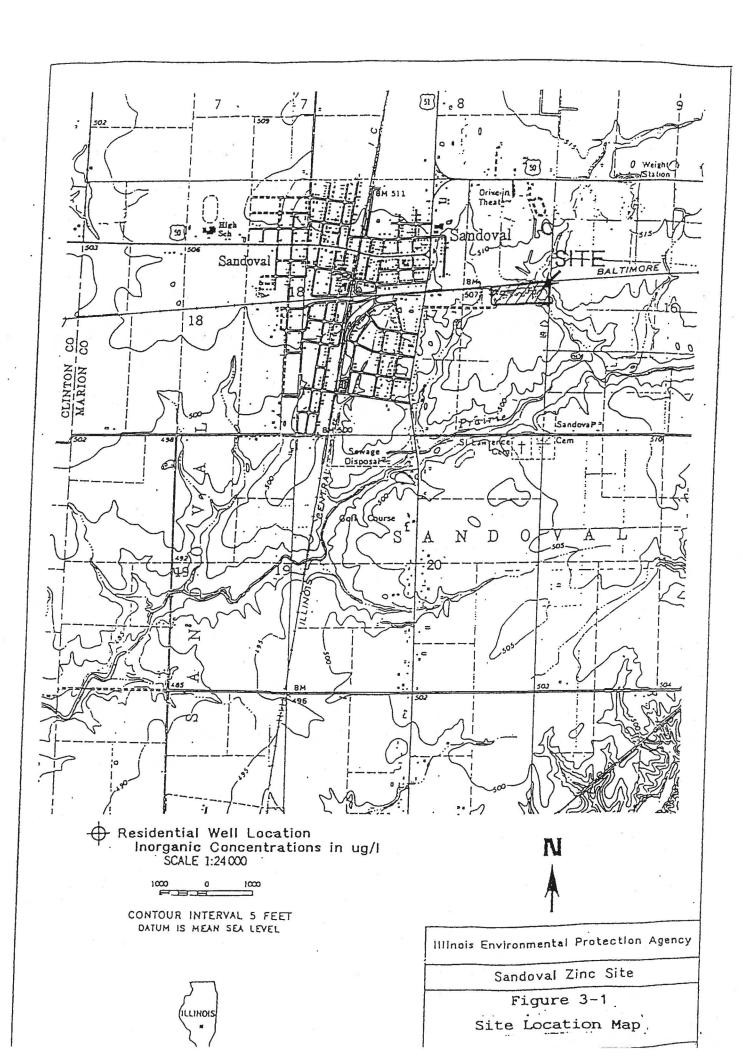
The hydraulic conductivity (permeability) of the Hagarstown Member was determined from the slug tests conducted at the newly installed EBASCO monitoring wells (MW01 and MW02) and at the IEPA monitoring wells (MW101, MW102, MW103)

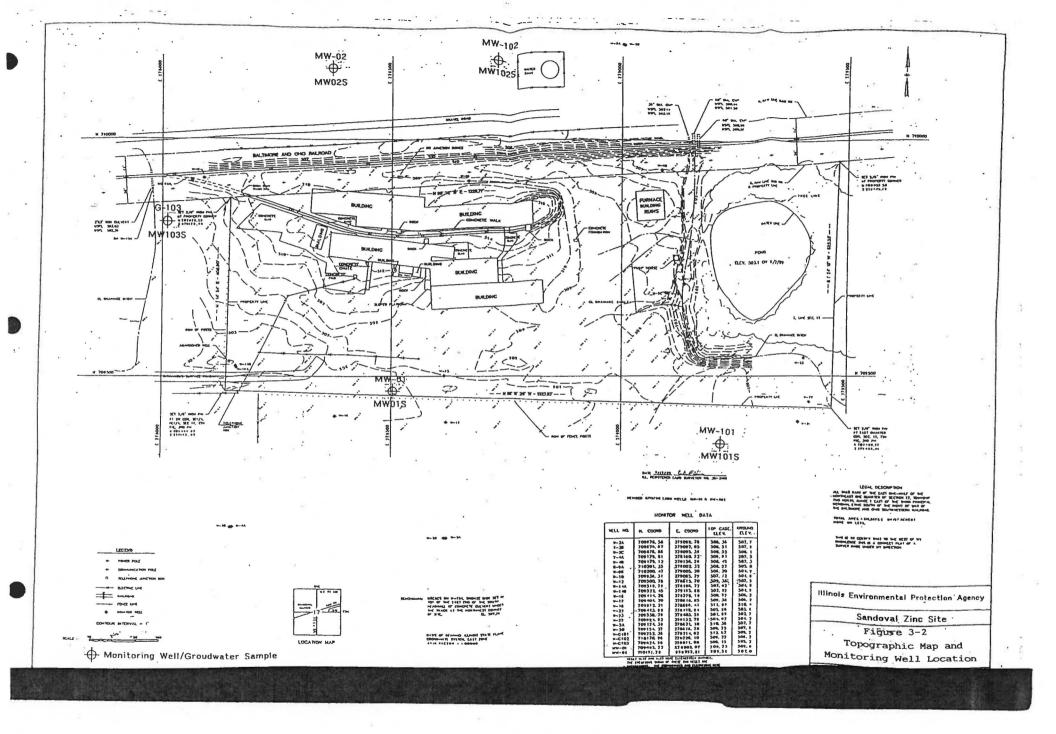
Slug tests consisting of falling and rising head tests were conducted on the wells in May and June 1990. Falling head slug tests were performed by instantaneously lowering a 4-1/2 foot long, 1-1/4 inch O.D. stainless steel slug attached to a nylon rope into the monitoring well until it was fully submerged. The water level drop was measured at timed intervals and recorded using a pressure transducer and data logger. Rising head slug tests consisted of quickly pulling the slug out of the well and recording the subsequent water level rise.

The hydraulic conductivity of the Hagarstown Member in the vicinity of the screened interval was calculated using the Hvorslev method for confined conditions (Hvorslev, 1957)<sup>7</sup>. The calculated hydraulic conductivity of the unit ranged from 2.2 to 4.9 ft/day (7.8x10<sup>-4</sup> to 1.7x10<sup>-3</sup> cm/s). Previous values reported for the Hagarstown ranged from 8.3x10<sup>-3</sup> to 9.1x10<sup>-3</sup> cm/sec (ISWS/ISGS, 1982). These values of hydraulic conductivity are consistent with the wide range of values reported in the literature for unconsolidated silty to clean sand (Freeze and Cherry, 1979)<sup>8</sup>. The slug test data and the hydraulic conductivity calculations are presented in Table 3-2.

## 3.5.2.4 Groundwater Velocity

According to Darcy's law, groundwater velocity is a function of hydraulic conductivity and hydraulic gradient. Since the hydraulic gradient cannot be determined due to the uncertainty of the groundwater flow direction in the Hagarstown Member, at present, the groundwater velocity cannot be estimated.



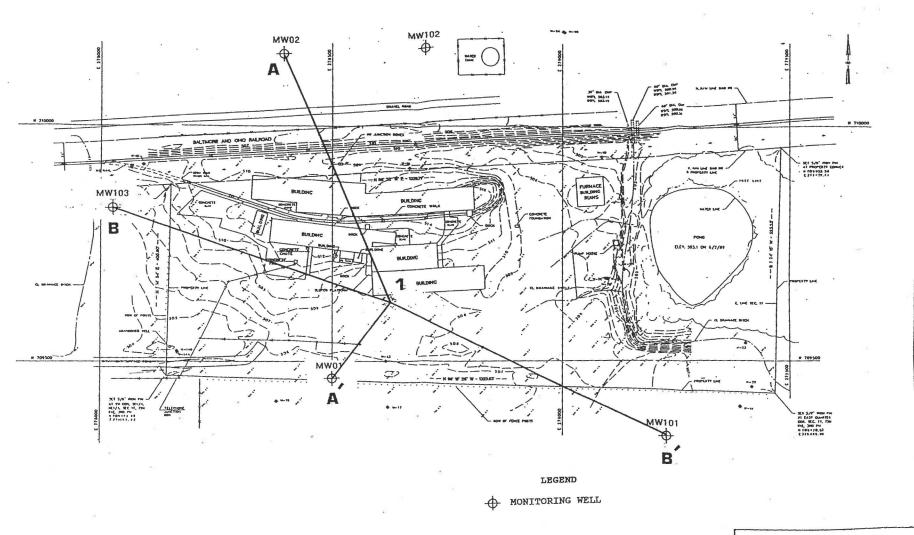


STAGE	UNIT
WISCONSINAN	PEORIA LOESS
WISCONSINAIN	ROXANA SILT
SANGAMONIAN	BERRY CLAY MEMBER- GLASFORD FORMATION
ILLINOIAN	HAGARSTOWN MEMBER- GLASFORD FORMATION
	GLASFORD FORMATION TILL
YARMOUTHIAN	LIERLE CLAY MEMBER- BANNER FORMATION
KANSAN	BANNER FORMATION TILL
PENSYLVANIAN SYSTEM	BOND FORMATION

(After Willman et al., 1975)

Sandoval Zinc Site

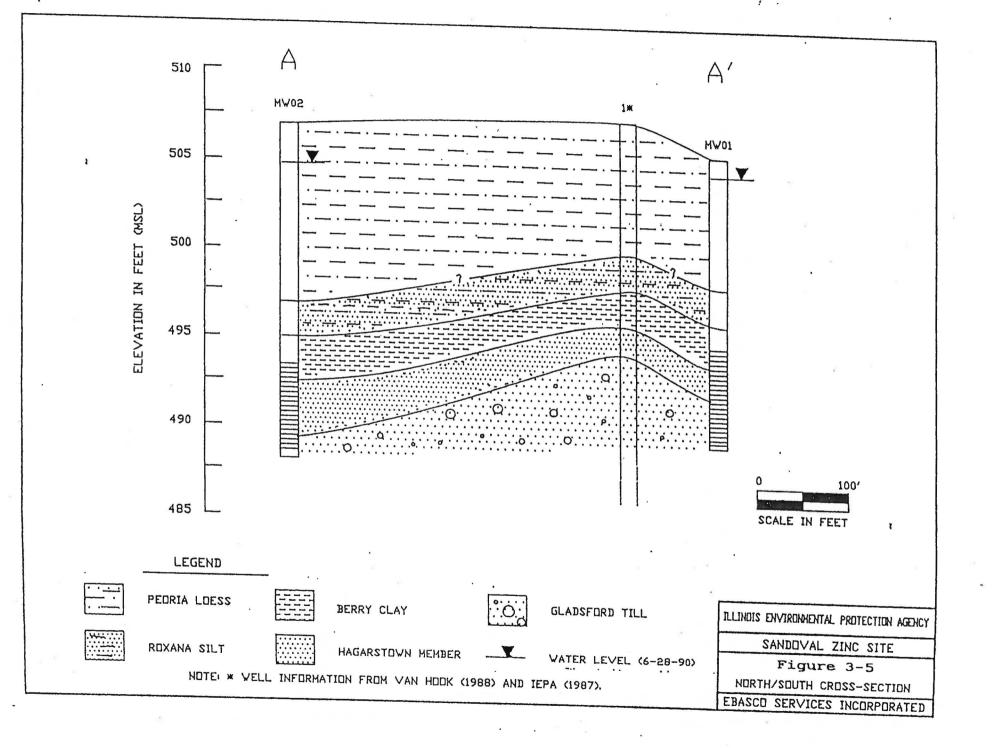
Figure 3-3
Generalized Stratigraphic

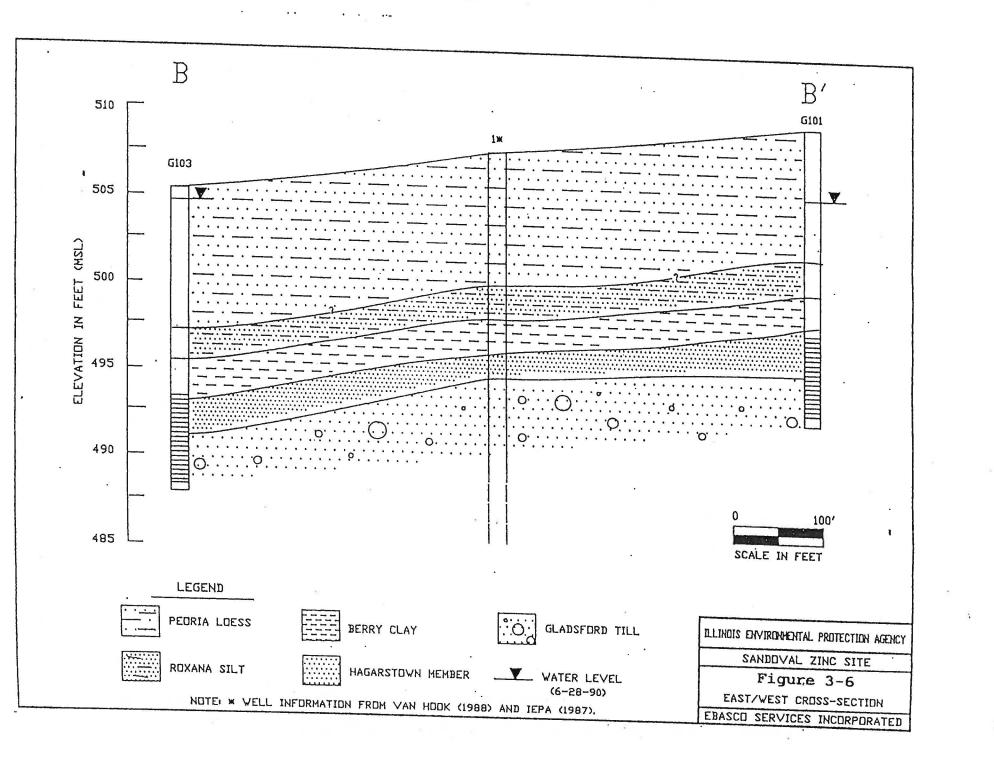


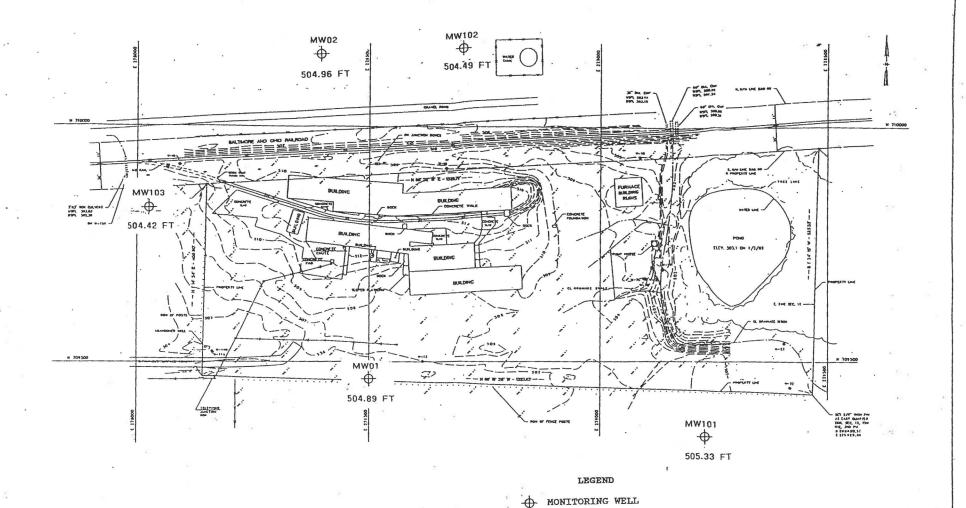
Illinois Environmental Protection Agency

Sandoval Zinc Site Figure 3-4

Generalized Geologic







Water level data collected on June 28, 1991.

1

Illinois Environmental Protection Agency

Sandoval Zinc Site

Figure 3-7
Monitoring Well
Water Level Elevations

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Kur.

Table 3-1 Monitoring Well Data Sandoval Zinc Sandoval, Illinois

WELL <sup>1</sup>	MWt01	MW102	MW to3	MW01	MW02
Elevation of Ground Surface (ft)	509.5	506.3	505.3	505.8	507
Elevation of Top of Casing (ft)	512.67	509:23	508.15	509.33	509.33
Depth to Top of Formation (ft)	11.2	12.2	12.3	12.5	14.4
Depth to Water (6/28/90) (ft)	7.34	4.74	4.74	4,44	4.42
Depth to Top of Screen (ft)	11.7	12	12	10.76	13.35
Elevation of Piezometric Surface (ft) (6/28/90)	505.33	504.49	504.42	504.89	504.96
Elevation of Top of Screen (ft)	497.8	494.3	493.3	495.04	493.65
Elevation of Top of Formation (ft) (Hagerstown)	498.3	494.1	493.3	495.04	492.6
Weii Completion Data	4/06/87	4/08/87	4/06/87	6/14/90	6/15/90

#### 1 - 5 foot screen

Note: IEPA monitoring wells G101, G102, and G103 were renamed for use in this report to MW101, MW102, and Mw103, respectively.

Table 3-2 Slug Test Analyses and Hydraulic Conductivity Calculations

Confined Conditions K=A/FT=d\*d\*In(2L/D+(1+2L/D)\*(2L/D))^0.5)/(8\*L\*T)

Well Number		Riser Diameter (D) (inches)	Borehole Diameter (D) (inches)	Length of Water Intake (L) (ft)	Basic Lag Time (T) (min)	Hydraulic Conductivity (K) (ft/day)
	* MW101	2	10.25	1.3	2.5	2.80E+00
	* MW102	2	10.25	2.7	2.1	2.23E+00
	* MW103	2	10.25	2	2.3	2.42E+00
	MW01	2	10.25	2.5	1.5	3.26E+00
	MW02	2	10.25	3.6	0.8	4.87E+00

\* IEPA monitoring wells G101, G102, and G103 were renamed for use in this report to MW101, MW102, and MW103, respectively.

## 4.0 NATURE AND EXTENT OF CONTAMINATION

This section presents a discussion of the analytical results, by sample media, for all samples collected during the field investigation at the Sandoval Zinc site. Detailed analytical results for all the samples are presented in Appendix C.

## 4.1 Above Ground Storage Tank

One sample and a duplicate were collected from the abandoned above ground storage tank (Figure 4-1). The samples were analyzed for Target Analyte List (TAL) inorganics, PCBs, and the volatile organic compounds benzene, toluene, ethyl benzene, and xylene (BTEX). The samples were also analyzed for their heating value. Five inorganic and three organic compounds were detected in the tank sample and in the duplicate. Iron, lead, nickel, vanadium, and zinc were all detected at concentrations under 50 ppm. Toluene, ethyl benzene, and xylenes were also detected in the tank sample and duplicate. Toluene values were 4,400 ppb and 6,700 ppb in the sample and duplicate; ethyl benzene was detected at 20,000 ppb and 23,000 ppb, and xylene at 96,000 ppb and 92,000 ppb. No PCBs were detected in either the tank sample or the duplicate. The heating values of the sample and duplicate were 18,500 and 17,800 btu/lb., respectively. The analytical result are summarized in Table 4-1.

## 4.2 Product/Ash

Six samples and one duplicate were collected from the piles of waste product and ash within the main building at the site (Figure 4-2). The samples were analyzed for full TAL inorganics and for EP Toxicity. Aluminum, iron, lead, and zinc were detected in concentrations greater than 10,000 mg/kg in at least two of the samples. Aluminum concentrations were greater than 10,000 mg/kg in five of the six samples and in the duplicate. Zinc concentrations were greater than 200,000 mg/kg in all samples and duplicate, except for sample WPA06S, where the level of zinc was 27,000 mg/kg. Other metals detected in relatively high concentrations were chromium, copper, and nickel, but they were found only in a random scattering across the samples. EP Toxicity results from the samples of waste product and ash varied from sample to sample. The maximum concentration levels permitted in the extract from EP Toxicity tests were exceeded in all samples for barium, cadmium, chromium, and lead, but the highest concentrations were found in samples WPA02S, WPA03S, WPA05S, and WPA06S.

Table 4-2 summarizes the key results for the inorganic analysis and the EP Toxicity test. The concentration levels for the EP Toxicity test are the legal limits for leachable metals.

All samples failed to meet the specified levels for one or more metals. Therefore, the waste product and ash must be considered hazardous waste.

## 4.3 Surface Soil

Twenty-three surface soil samples and three duplicate samples were collected from the locations shown in Figure 4-3. All the samples were analyzed for full TAL inorganics. Two of the samples, SS08S and SS10S, and the duplicate SS10D were also analyzed for PCBs. Aluminum, calcium, copper, iron, lead, and zinc were found in high concentrations in most samples. Concentrations of aluminum were typically greater than 5,000 ppm in the surface soil samples. Iron, lead, and zinc levels were found to be greater than 10,000 mg/kg in most samples and copper concentrations were typically above 1,000 mg/kg. Other metals that were detected at elevated levels in several samples include antimony, cadmium, chromium, manganese, mercury, nickel, and silver. No PCBs were detected in SS08S, SS10S, or SS10D. Table 4-3 shows a summary of the key analytical results for the surface soil samples.

## 4.3.1 Comparison of Results with Previous Investigation

A publication entitled "Retention of Zinc, Cadmium, Copper, and Lead By Geologic Materials" prepared by the Illinois State Water Survey (ISWS) and the Illinois State Geological Survey (ISGS) documents an investigation conducted at the Sandoval Zinc site from 1974 to 1977. The purpose of this investigation was to define the vertical and horizontal migration patterns of zinc, cadmium, copper, and lead through the soil and shallow aquifer systems at Sandoval Zinc and one other secondary zinc smelting site.

During the present field investigation, lead, zinc, copper, and nickel were detected in high concentrations in surface soil samples. Cadmium and silver were also detected but in relatively lower concentrations. The ISWS/ISGS study did not analyze soil samples for silver and nickel. Figure 4-3 shows the location of the surface soil samples collected in the present study along with concentrations of lead, zinc, copper, silver, and nickel in the samples. For comparison, Table 4-4 indicates the approximate concentrations of lead, zinc, cadmium, and copper obtained from selected well and core samples in the ISWS/ISGS study. The locations of these samples are shown in Figure 4-4 and are approximate since the ISWS/ISGS report did not use surveyed site maps to show sample locations. The ISWS/ISGS study did not analyze all core samples for the same parameters. Consequently, different ISWS/ISGS samples are compared to the same samples from the EBASCO study for specific analytes in Tables 4-4 and 4-5. The purpose of comparison between ISWS/ISGS data on metals concentration in surface soil, and the data compiled by EBASCO is to determine if site conditions have changed significantly since the ISWS/ISGS study.

In general, the data on surface soil samples in the EBASCO study are in the same range with those obtained in the ISWS/ISGS study. However, the exact concentrations of the laboratory analysis for the metals in the previous study are unknown. Therefore, EBASCO cannot be certain as to whether or not contaminants have migrated from the surface. Furthermore, the previous study did not analyze for silver which is found in concentrations significantly higher than those found in the average soils (0.01-5 mg/kg) throughout the United States. The silver could have come from the zinc ores mined from southern Missouri and smelted at the facility.

### 4.4 Surface Water

Four surface water samples and one duplicate were collected from the locations shown in Figure 4-1. The samples were analyzed for volatile organic compounds (BTEX) and for full TAL inorganics. The only volatile organic compound detected was toluene, but since toluene was also found in the laboratory blank, the compound could be due to laboratory contamination. Inorganic analytes detected in the surface water samples at elevated concentrations include aluminum, cadmium, calcium, copper, iron, magnesium, manganese, nickel, silver, thallium, and zinc (Table 4-6).

## 4.5 Groundwater

Two shallow monitoring wells were installed on-site. The wells were screened in the Hagarstown Member. Groundwater elevations in the two newly installed monitoring wells and the three existing wells on-site were measured on June 28, 1990. Water level data are insufficient to draw a contour map due to the small differences in elevations between the monitoring wells. It appears that the piezometric surface of the groundwater in the Hagarstown Member is relatively flat, so no determination of the hydraulic gradient or the groundwater velocity at the site could be made. Slug tests performed on the wells during the field investigation yielded hydraulic conductivity values ranging from  $8.8 \times 10^{-4}$  to  $2.8 \times 10^{-3}$  cm/sec for the Hagarstown Member.

Six groundwater samples and two duplicate samples were collected from five wells on-site (Figure 4-1) and from a single residential well (Figure 1-1, Residential Well A). Residential wells B&C were not sampled because they could not be identified. All samples were analyzed for full TAL inorganics; the samples collected from the five monitoring wells on-site were also analyzed for PCBs and BTEX. Two of the groundwater samples, MW102S and MW103S, were also analyzed for the full TCL organics list. None of the groundwater samples contained PCBs, nor did they contain any volatile organic compounds, with two

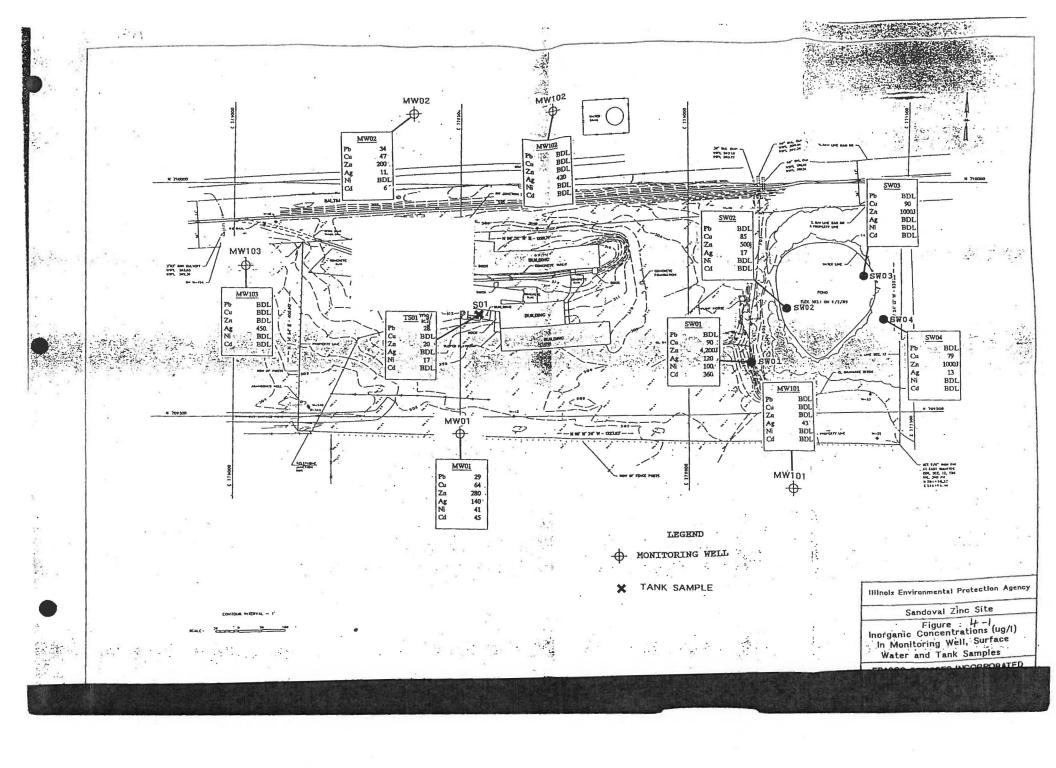
exceptions (Table 4-7). Groundwater samples MW01S and MW103S both contained trace amounts (less than  $5 \mu g/l$ ) of toluene.

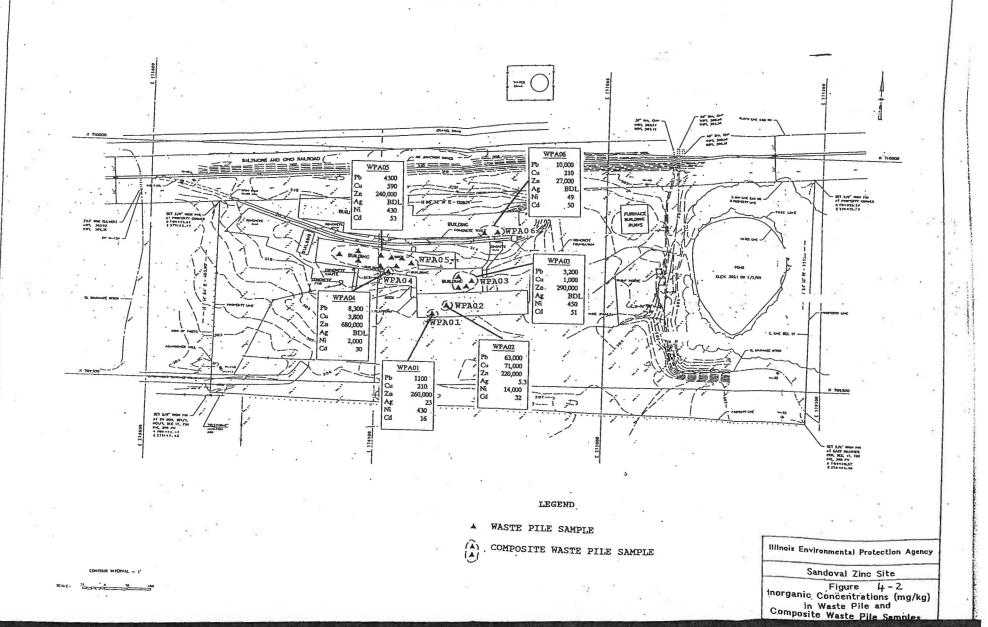
Of the twenty-three metals analyzed, calcium, magnesium, potassium, and sodium were detected at levels greater than 1,000  $\mu$ g/l in most samples. These concentrations are most likely due to the bedrock and soil composition and are probably unrelated to past site activities. Aluminum, cadmium, chromium, iron, manganese, silver, thallium, and zinc were also detected in most of the groundwater samples. The residential groundwater sample and the duplicate contained iron  $(2,660-2,700 \mu g/l)$ , manganese  $(160 \mu g/l)$  silver  $(60-61 \mu g/l)$ , thallium (100  $\mu$ g/l) and zinc 88-96  $\mu$ g/l). Silver and thallium in the residential well samples exceeded Federal Drinking Water Standards values of 50  $\mu$ g/l and 0.5  $\mu$ g/l, respectively. The groundwater samples collected from MW01 and MW02 exceeded Federal Drinking Water Standards for cadmium, chromium, copper, and silver. Table 4-8 summarizes the groundwater quality data. The high values of calcium (24,000-1,100,000  $\mu$ g/l) and magnesium  $(8,370-360,000 \mu g/1)$  indicate that these constituents were most likely released from the soil into the groundwater through ion-exchange with the contaminant metals onsite. So long as the soil has adequate ion-exchange capacity, the calcium and magnesium levels in the groundwater are likely to remain high.

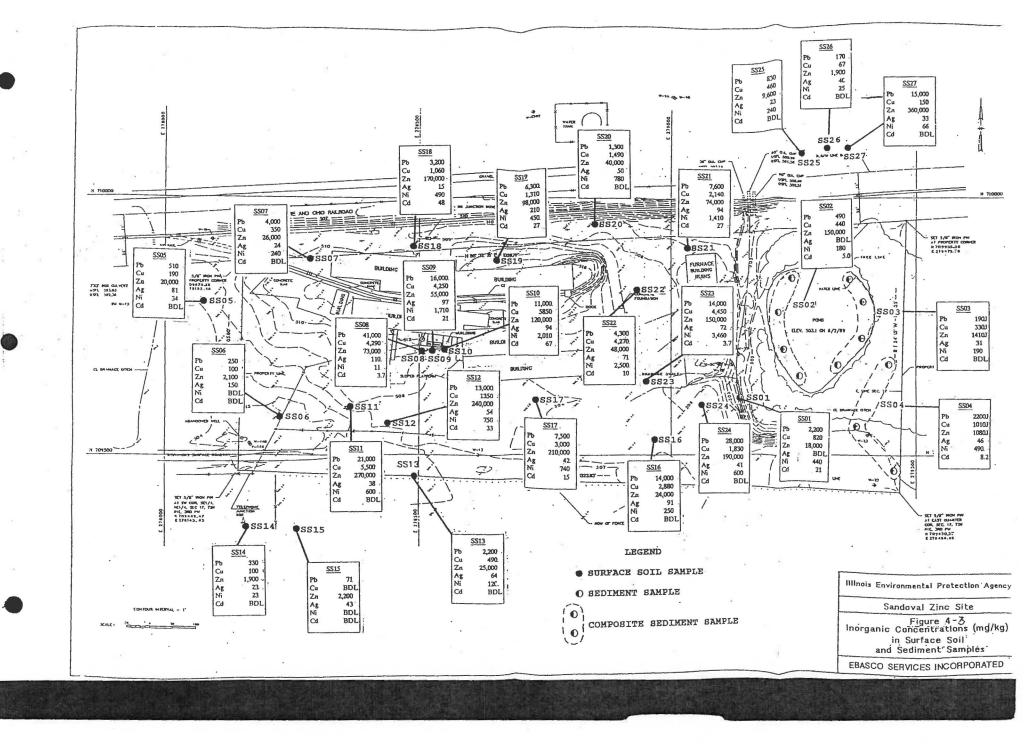
#### 4.6 Sediment

Sediment samples (SS01 through SS04) were collected from the four locations shown in Figure 4-3. Four samples and one duplicate were collected and analyzed for TAL inorganics and PCBs. High concentrations of aluminum, cadmium, copper, iron, lead, manganese, nickel, silver, and zinc were detected in all the sediment samples. Aluminum, iron, and zinc were found in concentrations generally greater than 10,000 mg/kg. The remaining detected metals were generally in the greatest concentrations in sediment samples SS01S, SS01D, and SS04S. No PCBs were detected in any of the sediment samples. Table 4-9 summarizes the key results from the analyses.

Although the sediment samples were not analyzed for EP Toxicity, the high levels of lead detected in the samples suggest that the sediments would be classified as a characteristic hazardous waste.







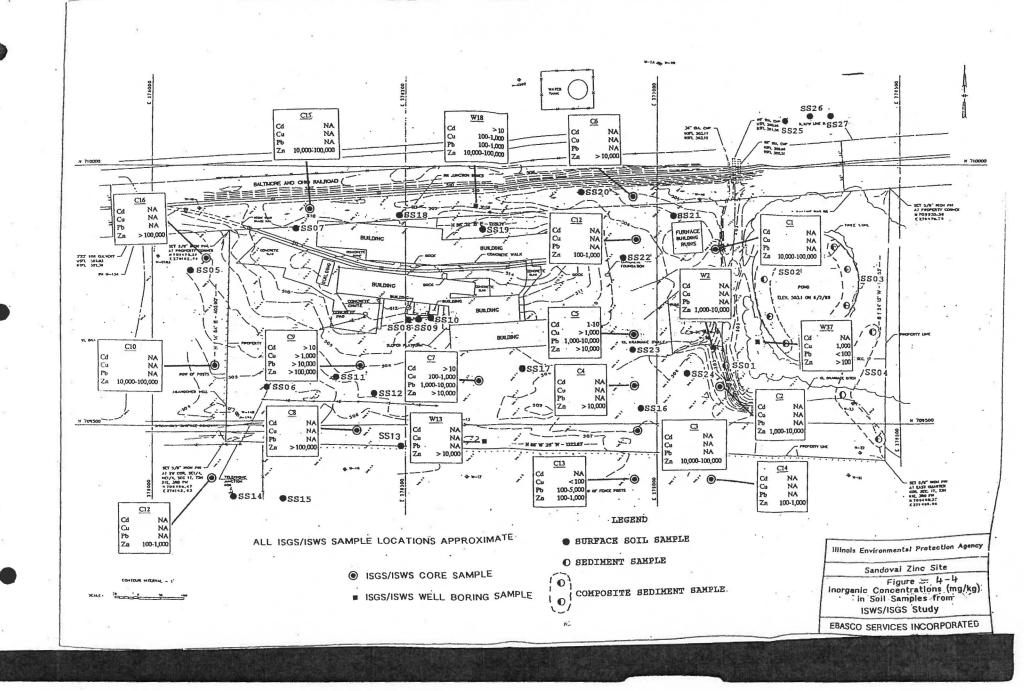


TABLE 4-1: SUMMARY OF ANALYTICAL RESULTS FROM TANK SAMPLES

	Sampling Locations								
Analysis	CRDL	TS01S	TS01D						
Inorganic Compounds (mg/kg)									
Iron	100	41	34						
Lead	5		28						
Nickel	40	17	17						
Vanadium	50	49	46						
Zinc	20	20	19						
Volatile Organics (µg/kg)									
Toluene	1.0	4400.J	6700						
Ethylbenzene	5.0	20,000.	23,000						
Xylenes (Total)	10.0	96,000.	92,000						
Heating Value (BTU/lb)		18,500	17,800						

S = Sample

D = Duplicate

J = Estimated Value

# TABLE 4-2: SUMMARY OF ANALYTICAL RESULTS FROM WASTE PRODUCT AND ASH SAMPLES

				Sampling L	ocations			
Analysis	CRDL	WPA01S	WPA01D	WPA02S	WPA03S	WPA04S	WPA05S	WPA06S
Inorganic Compounds (mg/kg)								
Aluminum Chromium Copper Iron Lead Nickel Zinc	200.0 10 25 100 5 40 20	74,000 330 210 87,000 1,100 430 260,000	79,000 300 160 94,000 1,000 450 260,000	37,000 330 71,000 22,000 63,000 14,000 220,000	27,000 110 1,000 24,000 3,200 450 290,000	10,000 40 3,800 2,300 8,300 2,000 680,000	15,000 55 590 62,000 4,300 430 240,000	1,800 1.9U 210 3,200 10,000 49 27,000
EP Toxicity Values (μg/l)	MCL	ś			140			
Barium Cadmium Chromium Lead	100.0 1.0 5.0 5.0	260 250 10 12	290 270 8.6 43	4,000 200 8.7 4,000	1,000 880 6.9 8,400	160 340 12 22,000	760 1,500 7.2 46,000	1,200 210 6.2 7,100

CRDL = Contract Required Detection Limits

MCL = Maximum Concentration Levels

S = Sample

D = Duplicate

TABLE 4-3: SUMMARY OF INORGANIC ANALYSES (MG/KG)
IN SURFACE SOIL SAMPLES

		Sampling Locations													
Analyte	CRDL	SS05S	SS05D	SS06S	SS07S	SS08S	SS09S	SS10S	SS10D	SS11S					
Aluminum	200.0	11,300	11,000	7,520	6,000	6,770	8,990	7,160	7,400	11,500					
Antimony Cadmium	60	2.8U	2.20	15U	15	28	16	23	17	61					
Calcium	5 000	1.4U	1.3U	1.2U	1.1U	3.7	21	67	35	1.50					
Chromium	5,000	10,900	3,500	96,800	14,300	1,830	3,430	4,570	2,440	1,640					
	10	2.8U	5.1	2.4U	24	9.5	16	98	18	60					
Copper	25	190	73	100	350	4,290	4,250	5,850	3,770	5,500					
Iron	100	36,300	19,700	14,900	18,500	75,200	69,700	70,300	58,900	26,900					
Lead	5	510	130	250	4,000	41,000	16,000	11,000	6,200	21,000					
Manganese	15	240	190	260	220	94	170	380	350	180					
Mercury	0.2	0.47	0.10U	0.11	1.0	0.22	0.45	0.21	0.19	2.1					
Nickel	40	34	44	9.7U	240	11	1,710	2,010	2,710						
Silver	10	81	47	150	24	110	97	94	83	600					
Zinc	20	20,000	4,200	2,100	26,000	73,000	55,000	120,000	88,000	270,000					

S = Sample

D = Duplicate

R = Rinsate

J = Estimated Value

TABLE 4-3: SUMMARY OF INORGANIC ANALYSES (MG/KG) (Cont'd.) IN SURFACE SOIL SAMPLES

		Sampling Locations													
Analyte	CRDL.	SS12S	SS12R	SS13S	SS13D	SS14S	SS14R	SS15S	SS16S	SS16R	SS17S				
41	200.0	11 000	00011	10,000	10.200	0.500	20011	44.400	10.400	00011	0.040				
Aluminum	200.0	11,800	200U	12,900	10,300	8,600	200U	11,100	10,400	200U	8,810				
Antimony	60	19	60U	13U	2.4U	2.6U	60U	2.2U	150	60U	25				
Cadmium	5	33	5U	1.10	1.2U	1.00	5U	1.20	1.0U	5U	15				
Calcium	5,000	1,800	1,000UJ	2,030	990	980	1,000UJ	630	5,210	1,000UJ	1,580				
Chromium	10	27	15	2.2	2.3U	2.0U	10	2.4U	24	10U	17				
Copper	25	1,350	53	490	520	100	35	6.0U	2,880	25	3,000				
Iron	100	35,300	66J	36,500	42,100	15,900	50UJ	22,300	57,600	64J	37,400				
Lead	5	13,000	52	2,200	2,200	330	5U	71	14,000	37	7,500				
Manganese	15	550	15U	1,150	1,360	400	15U	2,320	410	15U	400				
Mercury	0.2	49	0.20U	0.37	1.1	0.069U	0.20U	0.059U	0.37	0.20U	13				
Nickel	40	750	40U	120	110	23	40U	9.5U	250	40U	740				
Silver	10	54	10U	64	67	23	10U	43	91	10U	42				
Zinc	20	240,000	3,200	25,000	21,000	1,900	440	2,200	24,000	100	210,000				

S = Sample

D = Duplicate

R = Rinsate

J = Estimated Value

TABLE 4-3: SUMMARY OF INORGANIC ANALYSES (MG/KG) (Cont'd.) IN SURFACE SOIL SAMPLES

Analyte		Sampling Locations												
	CRDL	SS18S	SS19S	SS20S	SS21S	SS22S	SS23S	SS24S	SS25S	SS26S	SS27S			
Aluminum	200.0	6,130	6,530	7,770	10,700	7,310	6,540	5,710	8,910	9,630	C 750			
Antimony	60	65	240	12U	280	210	180	60	6.6	2.6U	6,750			
Cadmium	5	48	27	1.0U	27	10	3.7	1.3U	1.4U	1.4U	2.40			
Calcium	5,000	4,180	23,500	1,670	29,200	2,090	4,500	750	960		1.00			
Chromium	10	13	73	4.4	8.3	14	16	2.6U	2.7U	480	1,270			
Copper	25	1,060	1,310	1,490	2,140	4,270	4,450	1,830	460	2.70	2.00			
Iron	100	5,380	126,000	32,100	56,600	44,600	54,100	43,200	16,300	67	150			
Lead	5	3,200	6,300	1,300	7,600	4,300	14,000	28,000	830	21,000	18,400			
Manganese	15	340	3.5U	4.2	320	120	290	13		170	15,000			
Mercury	0.2	1.4	5.7	0.45	0.66	0.53	0.46	7.7	390	1,790	910			
Nickel	40	490	450	780	1,410	2,500	3,460	600	0.11	0.098U	0.081U			
Silver	10	15	210	50	94	71	72	25 (2) 27	240	25	66			
Zinc	20	170,000	98,000	40,000	74,000	48,000	150,000	41 190,000	9,600	40 1,900	33 360,000			

S = Sample

D = Duplicate

R = Rinsate

J = Estimated Value

TABLE 4-4: SUMMARY OF SURFACE SOIL ANALYTICAL DATA (All Concentrations in mg/kg)

	LE	AD			ZII	VC		COPPER				
Ebasc	o Study	ISGS	Study	Ebaso	o Study	ISGS	ISGS Study		Ebasco Study		Study	
Sample #	Concentration	Sample #	Concentration	Sample #	Concentration	Sample #	Concentration	Sample #	Concentration	Sample #	Concentration	
Background:	71-15,000	Background:	10-40	Background:	1,900-360,000	Background:	20-50	Background:	6-460	Background:		
SS14 SS15	330 71	# ·		SS14 SS15	1,900 2,200	C12	100-1,000	SS14 SS15	100 6.0U	8		
SS25 SS26 SS27	830 170 15,000			SS25 SS26 SS27	9,600 1,900			SS25 SS26	460 67		w. K	
SS01	2,200	C37	<100		360,000	W2	1,000-10,000	SS27 SS01	150 820	C37	1 000	
SS02* SS03* SS04*	190J	× 1		SS02* SS03*	150,000 1,410J	C1	10,000-100,000	SS02* SS03*	440 330J	Gur	1,000	
SS05 SS06	2,200J 510 250	wз	100-1,000	SS04* SS05	1,080J 20,000	C6	10,000	SS04* SS05	1,010J 190	w3	<100	
SS07 SS08	4,000 41,000			SS06 SS07 SS08	2,100 26,000	C15	10,000-100,000		100 350			
SS09 SS10	16,000 11,000			SS09 SS10	73,000 55,000 120,000			SS08 SS09	4,290 4,250			
SS11 SS12	21,000 13,000	C9	>10,000		270,000 270,000 240,000	C19	>100,000	SS10 SS11 SS12	5,850 5,500 1,350	C9	>1,000	

<sup>·</sup> Composite of 3 Grab Samples

## TABLE 4-4: SUMMARY OF SURFACE SOIL ANALYTICAL DATA (Cont'd.) (All Concentrations in mg/kg)

	LEA	\D			ZIN	IC			COP	PER	
Ebasc	Ebasco Study ISGS Study		Study	Ebasco Study ISGS			Study Ebasco S		o Study	ISGS	Study
Sample #	Concentration	Sample #	Concentration	Sample #	Concentration	Sample #	Concentration	Sample #	Concentration	Sample #	Concentration
SS13	2,200	C13	100-5,000	S <b>S</b> 13	25,000	W13	>10,000	SS13	490	C13	<100
SS16	14,000			SS16	24,000	C8	>100,000	SS16	2,880		
SS17	7,500	C7	1,000-10,000	SS17	210,000	C4	10,000-100,000	SS17	3,000	C7	100-1,000
SS18	3,200			SS18	170,000	C7	>10,000	SS18	1,060		
SS19	6,300	W18	100-1,000	SS19	98,000	W18	10,000-100,000	SS19	1,310	W18	100-1,000
SS20	1,300	Æ		SS20	40,000	C6	10,000-100,000	SS20	1,490		
SS21	7,600		= ==	SS21	74,000	C10, C1	10,000-100,000	SS21	2,140		
SS22	4,300			SS22	48,000	C12	10,000-100,000	SS22	4,270		
SS23	14,000	C5	1,000-10,000	SS23	150,000	C5	10,000-100,000	SS23	4,450	C5	>1,000
SS24	28,000			SS24	190,000	C2	1,000-10,000	SS24	1,830		

TABLE 4-5: SUMMARY OF SURFACE SOIL ANALYTICAL DATA (All Concentrations in mg/kg)

	CADI	MUM		NIC	CKEL	SILVER		
Ebaso	co Study	ISGS Study		Ebasco Study		Ebasco Study		
Sample #	Concentration	Sample #	Concentration	Sample #	Concentration	<del></del>	Concentration	
Background:	1.0-1.5U		0.04-1.5	Background:	0.070-240	Background:	23-40	
SS14 SS15	1.0U 1.2U			SS14		SS14	23	
SS25	1.4U			SS15 S <b>S2</b> 5		SS15	43	
SS26	1.4U			SS26		SS25	23	
SS27	1.0U			SS27		SS26 SS27	40 33	
SS01	21			SS01	440	SS01	0.01	
SS02	5.0			SS02		SS02	3.20	
SS03	1.5U			SS03		SS03	3.6U 31	
SS04	8.2			SS04		SS04	46	
SS05	1.4U	W3	1-10	SS05		SS05	81	
SS06	1.2U			SS06	9.7U	SS06	150	
SS07	1.10	i e		SS07	240	SS07	24	
SS08 SS09	3.7			SS08	.11	SS08	110	
SS10	21			SS09	1,710	SS09	97	
SS10	35-67*	C9	>10	SS10	(5)	SS10	94	
SS12	1.5U .			S <b>S</b> 11		SS11	38	
SS12	1.10			SS12		SS12	54	
SS16	1.00			SS13	1	SS13	64	
	1.00			S <b>S</b> 16	250	S <b>S</b> 16	91	

<sup>\*</sup> Two samples were collected including one duplicate sample. The concentrations were 35 and 67 mg/kg.

TABLE 4-5: SUMMARY OF SURFACE SOIL ANALYTICAL DATA (Cont'd.)
(All Concentrations in mg/kg)

	CADI	MUIM	n *ac	N	ICKEL	SILVER		
Eba	sco Study	ISGS Study		Ebasco Study		Ebasco Study		
Sample #	Concentration	Sample #	Concentration	Sample #	Concentration	Sample #	Concentration	
SS17	15	C7	>10	SS17	740	SS17	42	
SS18	48			SS18	490	SS18	15	
SS19	27	W18	>10	SS19	450	SS19	210	
SS20	1.0U			SS20	780	SS20	50	
SS21	27			SS21	1,410	SS21	94	
SS22	10	-		SS22	2,500	SS22	71	
SS23	3.7	C5	1-10	SS23	3,460	SS23	72	
SS24	1.3U			SS24	600	SS24	41	

U - Compound analyzed for but not detected. Value reported is Contract Required Detection Limit (CRDL).

## TABLE 4-6: SUMMARY OF ANALYTICAL DATA FOR SURFACE WATER SAMPLES

	Sampling Locations								
Analysis	CRDL	SW01S	SW01D	SW02S	SW03S	SW04S	SW04R	Federal Drinking Water Standards MCL	
Inorganic Compounds (mg/l)  Aluminum Cadmium Calcium Copper Iron Magnesium Manganese Nickel Silver Thallium Zinc	200.0 5 5,000 25 100 5,000 15 40 10 20	780 360 100,000 90 1,400 16,000 1,500 100 120 53 4,200J	1,000 370 110,000 80 1,400 16,000 1,600 100 120 47 4,100J	5,200 5.0U 17,000 85 3,200 2,400 84 40U 17 10U 500J	5,600 5.0U 5,300 90 3,300 2,500 120 40U 10U 10U 1,000J	660 5.0U 18,000 79 3,200 4,400 930 40U 13 10U 1,000J	200U 5.0U 1,000U 25U 50U 1,000U 15U 40U 10U 10U	0.005  1.3   0.10	(Proposed (Proposed (Tentative)
<i>Volatile Compounds (µg/I)</i> Toluene	1.0	5U	5U	5U	25B	5B	27B		

CRDL = Contract Required Detection Limits

S = Sample

D = Duplicate

R = Rinsate Sample

J = Estimated Value

TABLE 4-7: SUMMARY OF ANALYTICAL DATA FOR GROUNDWATER SAMPLES

	Sampling Locations									
Analysis	CRDL	MW101S	MW101D	MW101R	MW102S	MW103S	MW01S Total	MW02s Total	RW01S	RW01D
Inorganic Compounds (mg/l)										
Aluminum Cadmium Calcium Chromium Copper Iron Magnesium Manganese Potassium Silver Sodium Thallium Zinc  Volatile Compounds (µq/l)	200.0 5 5,000 10 25 100 5,000 15 5,000 10 5,000	200U 5.0U 24,000 10U 25U 50U 13,000 15U 670 43 240,000 10U 20U	200U 5.0U 24,000 10U 25U 50U 13,000 15U 670 40 241,000 10U 20U	200U 5.0U 1,000U 10U 25U 50U 1,000U 15U 500U 10U 500U 10U 20U	200U 5.0U 240,000J 10U 25U 50U 170,000 15U 1,400 420 280,000 180 20U	200U 5.0U 290,000J 10U 25U 50U 150,000 380 3,000 450 95,000 190 20U	14,000J 45 1,100,000 150 64 34,000J 360,000J 1,500J 6,500 140 420,000 5.0U 280	13,000J 6.0 130,000 46,000J 46,000J 1,400J 5,900 11 230,000 5.0U 200	200U 5U 66,900J 10U 35 2,700J 8,370 160 1,240 66 12,300 10U 88	200U 5U 68,800J 10U 2,660J 8,400 160 1,230 61 12,400 110 96
Toluene	1.0	5U	5U	5U	5U	2J	. 4J	<b>5</b> U	NR	NF

CRDL = Contract Required Detection Limits

S = Sample

D = Duplicate

R = Rinsate

J = Estimated Value

U = Compound Analyzed for But Not Detected

NR = Analysis Not Run

RW = Residential Well

# TABLE 4-8: SUMMARY OF GROUNDWATER QUALITY DATA FOR THE SANDOVAL ZINC SITE

	Range of Values for	Federal Drinking Water Standards As of April, 1990				
Analyte	Monitoring and Residential Well Water Samples  µg/l	NIPDWR (1) µg/l	MCL (2)/MCLG (3) μg/l	MCLS (4) mg/l		
Aluminum	ND (5) - 14,000			0.05 to 0.2		
Cadmium	6–45	10	5 (P)/5 (P)			
Chromium	69–150	50	100 (P)/100 (P)			
Copper	35–64		1,300 (P)/1,300 (P)	1		
Iron	ND (5) - 34,000			0.3		
Manganese	ND (5) – 1,500			0.05		
_ead	29–34		5 (P)/Zero (P)	*		
Nickel	41		100 (T)/100 (T)	3		
Silver	11–450	50		0.09		
Toluene	ND (5) – 4	·	2,000 (P)/2,000 (P)	0.04		
Zinc	88–280			5		
Calcium	24,000-1,100,000					
Magnesium	8,370–360,000					
		· ·				

- (1) National Primary Drinking Water Standard
- (2) Maximum Contaminant Level
- (3) Maximum Contaminant Level Goal
- (4) Secondary Maximum Contaminant Levels
- (5) None Detected
- (P) Proposed Regulatory Value
- (T) Tentative Regulatory Value

TABLE 4-9: SUMMARY OF ANALYTICAL RESULTS FROM SEDIMENT SAMPLES

	Sampling Locations						
Analysis	CRDL	SS01S	SS01D	SS02S	SS03S	SS04S	SS04R
Inorganic Compounds (mg/kg)		,					
Aluminum	200.0	18,000	12,000	9,600	8,560	13,600	200U
Cadmium	5	21	19	5.0	1.5U	8.2	5.0U
Copper	25	820	850	440	330J	1,010J	87
Iron	100	15,000	12,000	13,000	17,100J	66,400J	24,100J
Lead	5	2,200	2,000	490	190J	2,200J	120
Manganese	15	260	1,300	290	270	2,770	170
Nickel	40	440	470	180	190	490	198
Silver	10	3.2U	2.9U	3.6U	31	46	10U
Zinc	20	18,000	15,000	150,000	1,410J	1,080J	16,200

CRDL = Contract Required Detection Limits

S = Sample

D = Duplicate

R = Rinsate

J = Estimated Value

U = Compound Analyzed for But Not Detected

#### 5.0 SUMMARY AND CONCLUSIONS

This section summarizes the major findings of the field investigation. The primary contaminants of concern are associated with the past operation and maintenance of the Sandoval Zinc Company. These contaminants are primarily the heavy metals from the smelting process. A summary of the extent of the inorganic contamination in the soils, groundwater, and waste products is presented.

Analytical results of the tank sample and the duplicate show that the tank contains residual fuel oil with an average heating value of 18,100 BTU/lb. The oil does not contain PCBs but contains 28 mg/kg of lead. Other inorganic analytes detected in low concentrations (less than 50 ppm) were iron, nickel, vanadium, and zinc. Because of the high lead concentration, the oil would be classified as a characteristic hazardous waste (D008).

The ash and waste product inside the buildings contain high concentrations of aluminum, iron, lead, and zinc. Zinc concentrations were typically greater than 200,000 mg/kg. All of the samples collected failed the EP Toxicity test for barium, cadmium, chromium, and lead. However, the waste product and ash on-site are not listed hazardous wastes, but would be classified as characteristic hazardous waste.

The inorganic analytes detected in the surface soil samples include copper, lead, nickel, and zinc. Zinc concentrations ranged from 1,900 mg/kg to 360,000 mg/kg in the samples. Copper, lead, and nickel concentrations typically were much lower, from 10 to 50,000 mg/kg. These concentrations correlate reasonably well with surface soil data from the previous ISWS/ISGS study. Based on the site geology and close correlation of surface concentrations, data on the subsurface soil conditions from the previous study should still be valid and representative of subsurface conditions at the site. This assumption is reasonable in light of the low permeability of the underlying till material. Consequently, the ISWS/ISGS data can be used to approximate volumes of contaminated on-site soils for the feasibility study.

Hydraulic conductivity values obtained from EBASCO's field investigation ranged from  $8.8 \times 10^{-4}$  to  $2.8 \times 10^{-3}$  cm/sec and are within the normal range for silty sand. However, no determination can be made regarding the hydraulic gradient or the groundwater velocity.

No PCBs were found in the groundwater samples, but they do contain high concentrations of cadmium, chromium, copper, and silver. These contaminants could have been transported from the impacted surface soil to the groundwater via abandoned investigative wells which have not been plugged and/or damaged and improperly installed wells that still exist on-site. The groundwater samples collected from MW01 and MW02, the newly

installed wells, both exceed the National Interim Primary Drinking Water Regulations (NIPDWR). The residential well sample contained levels of silver and thallium exceeding the NIPDWR standards.

There is considerable difference in the water quality data between the two newly installed wells and the three old monitoring wells. Additional sampling would be required to resolve this discrepancy and establish whether or not samples from the new wells are representative of current site conditions. Filtered samples show concentrations of dissolved metals and are more important from the stand point of compliance with drinking water standards, unfiltered samples represent a worst case scenario for determining treatment options.

Additional monitoring wells will probably not be necessary to characterize the groundwater quality on-site. A conventional pump and treat system is not likely to be considered a favorable alternative to remediate the groundwater at this facility because of the low productivity of the Hagarstown formation. There is no immediate health concerns for the drinking water at Sandoval because the city receives drinking water from Centralia. The current groundwater quality on-site poses no threat to the deep productive water bearing aquifer in the Hagarstown formation so long as the soil has adequate ion-exchange capacity and the contaminant metals are retained in the soil. However, the potential for groundwater transport of metals can be substantially reduced or eliminated by removing the metals from the soil or immobilizing them in the soil.

Based on one round of sampling, the concentrations of cadmium, copper, nickel, and silver in the surface water from the farm pond exceed the MCLs for drinking water. This water would require treatment prior to discharge.

Sediments in the vicinity of the farm pond also contain high concentrations of metals. The farm pond has not been previously investigated, and additional sampling to further define the extent of contamination for remediation would be required as part of the remedial design for this site.

The high levels of lead and zinc in the sediments suggest that these metals were probably not transported to the farm pond area through surface water runoff or groundwater movement. The terrain on-site is essentially flat and is not conducive to such transport. The high metals concentration in the sediments could have resulted from using the farm pond as a processing unit to store waste water when the smelters were in operation. These sediments would likely be classified as characteristic hazardous waste.

#### 6.0 IDENTIFICATION AND SCREENING OF TECHNOLOGIES

This FS was performed according to the following steps:

- Establish potential remedial objectives.
- Identify general response actions to meet remedial objectives, including no action.
- Identify remedial technologies under each general response action with emphasis on permanent solutions.
- Screen remedial technologies based on technical considerations and then, use those technologies to develop remedial alternatives.
- Screen remedial alternatives according to effectiveness, implementability, and cost.
- Perform a detailed evaluation of the remaining remedial alternatives based on short-term effectiveness; long-term effectiveness and permanence; reduction of toxicity, mobility, and volume; implementability; cost; compliance with ARARs; overall protection of human health and the environment; and state and community acceptance, and
- Perform a comparative evaluation between remedial alternatives.

The FS methodology for each of these steps is described in detail in the appropriate sections.

This section summarizes the screening process used to identify technologies appropriate to remedy contaminants of concern at the Sandoval Zinc site.

#### 6.1 Remedial Action Objectives

The IEPA established Remedial Action Objectives (RAOs) based on the results of the additional field investigation, the ISWS/ISGS study, and the concentration of contaminants considered to be acceptable for the site. These objectives are listed in Table 6-1 by parameter separately for groundwater and soil.

The IEPA's current remediation strategy is to identify and evaluate those remedial technologies and process options that can achieve the established RAOs. These are numerical objectives, which if attained, would allow the site to be restored for unrestricted use. These objectives do not take into consideration contaminant pathways, potential receptors and the potential carcinogenic and non-carcinogenic risk posed by the contaminants to those receptors.

For the purpose of this FS, EBASCO has assumed that prevention of exposure to the contaminants is also a remedial action objective. This will facilitate evaluation of those remedial technologies that cannot achieve the numerical objectives, but can be effective in eliminating the risk of exposure to contaminants.

# 6.2 The Study Area for the Feasibility Study

The study area for this Feasibility Study is shown in Figure 6-1. After EBASCO completed the additional field investigation in 1990, the IEPA installed a fence around the site to restrict public access as part of the initial step to prevent exposure to contaminants on-site. This fence line essentially outlines the boundary established by the IEPA for the purpose of this Feasibility Study. However, also included within this boundary is the "farm pond" located east of the site. Any area outside the designated boundary is beyond the scope of this Feasibility Study.

# 6.3 Impacted Areas for the FS

This section summarizes the five areas of concern that are addressed in this FS. These areas are: 1) the above ground storage tank, 2) waste product/ash and miscellaneous debris, 3) impacted soil, 4) impacted groundwater, and 5) the farm pond and associated impacted sediment.

# 6.3.1 Aboveground Storage Tank

The aboveground storage tank (AST) located on-site was found to contained residual fuel oil. A sample of the fuel oil was collected for laboratory analysis (see Section 4.0) and was determined to have sufficient heating value to be used as supplementary fuel for combustion. However, in September 1991, a valve in the outlet line failed and released a considerable portion of the tank contents onto the ground. The IEPA implemented emergency response action to mitigate the immediate hazards posed by the spill. As a result of the emergency response action all of the visibly impacted soils have been removed and

are presently stored on a plastic liner inside one of the on-site buildings. The volume of impacted soil removed was approximately 500 yd.<sup>3</sup>. Both the impacted soil and the aboveground storage tank, which is presently empty, will require proper disposal.

# 6.3.2 Waste Product/Ash and Miscellaneous Debris

Approximately 5,000 lbs. of waste product/ash is present inside the on-site buildings. Based on the EP Toxicity Test results (see Section 4.0) this material is considered a characteristic hazardous waste. The miscellaneous debris on-site consist of building rubble, remains from the old smelter, and other general debris inside the buildings. The volume of the miscellaneous debris is estimated to be 1,500 yd.<sup>3</sup>.

# 6.3.3 <u>Impacted Soil</u>

Field investigation results indicate that both the surface and the sub-surface soil are impacted with heavy metals. Based on the IEPA established clean up objectives presented in Table 6-1, more than 425,000 cubic yards of impacted soil requires remediation. An estimate of the volume of soil to be remediated is shown in Table 6-2. The calculations for the soil volume estimate are provided in Appendix D.

In estimating the volume of soil to be remediated, the required dimensions were taken from the surveyed map prepared as part of EBASCO's Additional Field Investigation, the report on the previous study conducted by ISWS/ISGS, and the sketch provided by the IEPA to delineate the site boundary for this FS. Based on the available information regarding metal concentrations in the soil, the site area to be remediated for each metal is estimated to be 510,000 ft<sup>2</sup> (1200 ft x 45 ft). The depth to which remediation would be required depends upon the specific metal and its concentration in the soil. For example, the cadmium concentration is greater than 1 mg/kg at depths up to 17.5 ft below the surface, whereas lead concentration is greater than 100 mg/kg at the same depth. The depth to specific concentrations were taken from the concentration profile charts presented in the ISWS/ISAS report. Sample calculations for estimated volumes of impacted soil are presented in Appendix D.

In the ISWS/ISGS study, sub-surface soil samples were collected up to a depth of approximately 28 feet. The concentrations listed in Table 6-2 are the lowest concentrations for which sub-surface analytical data is currently available. These data cannot be reliably extrapolated to determine the depths at which the IEPA established RAOs can be achieved because the concentration profile charts in the ISWS/ISAS report do not indicate any specific correlation between depth and soil concentration. Consequently, the table

represents the minimum volume of soil requiring remediation to achieve the concentrations listed.

# 6.3.4 Impacted Groundwater

To estimate the volume of groundwater requiring remediation, the groundwater in contact with the soil was assumed to cover an area equivalent to that covered by the soil (i.e., 510,000 ft<sup>2</sup>). Since contaminants were found at depths up to 28 feet, a total depth of 30 feet was assumed as the depth up to which groundwater in contact with the soil is expected to be impacted. Since the average depth to groundwater on-site is approximately 5 feet, the estimated thickness of the impacted water is 25 feet. The total volume occupied by impacted soil and associated groundwater would thus be a cube with dimensions of 1200 ft x 425 ft x 25 ft. Since the groundwater exists in the interconnected pores of the soil, only a portion of this cube volume can be attributed to the groundwater. A porosity of 15% was assumed for this calculation. In addition to this, some groundwater also exists as moisture in the soil above the water table. This is estimated to be 10% of the volume occupied by the soil.

Thus, the estimated volume of impacted groundwater beneath the site to be remediated is 15.7 X 10<sup>6</sup> gallons. Appendix D presents sample calculations to show how this volume was calculated.

### 6.3.5 Farm Pond and Associated Sediment

Field investigation results (Section 4.0) indicate that the both the surface water and sediments of the "farm pond" contain elevated levels of heavy metals. These metals include aluminum, cadmium, copper, iron, lead, nickel, silver and zinc. EP Toxicity test were not conducted on the sediment samples collected, however, aluminum, iron and zinc were all detected at concentrations greater than 10,000 mg/kg. Lead was also detected at concentrations of greater than 2,000 mg/kg. Therefore, it is likely that the sediments will possess hazardous waste characteristics.

To estimate the volume of impacted sediments requiring remediation, it was assumed that the metals were present in the top one foot of sediment. The "farm pond" is approximately one acre in size, therefore, an estimated 43,500 ft<sup>3</sup> (1,600 yd<sup>3</sup>) of impacted sediment is present on the bottom of the pond.

# 6.4 General Response Actions

This section presents general response actions identified to meet the RAOs established for the Sandoval Zinc site. Table 6-3 summarizes the general response actions which were determined to be feasible for the site.

These general response actions (GRAs) were selected from a comprehensive list of general response actions typically considered for the clean-up of hazardous waste sites. The selections were based on information obtained from the Additional Field Investigation, the ISWS/ISGS study (1982), and site specific conditions. The GRAs were developed from the October, 1988 Interim Final RI/FS Guidance Document (USEPA, 1988), The Superfund Innovative Technology Evaluation Program: Technology Profiles (USEPA, 1990), information obtained from the Alternative Treatment Technology Information Center (ATTIC), experience on other hazardous waste projects, knowledge of new technologies, and the professional judgment of the engineers performing the Feasibility Study. For example, remedial technologies designed to remove or destroy organic contaminants were not considered since heavy metals are the primary contaminants of concern at the site.

# 6.5 Identification and Screening of Technology

The next step in the screening process is to identify the remedial technologies associated with each general response action applicable to the Sandoval Zinc site and then to determine their feasibility. Each applicable general response action contains many remedial technologies, and an exhaustive list could be developed from various USEPA guidance documents and handbooks, as well as from other feasibility studies. However, some of these technologies are obviously not applicable to this site. Therefore, this identification and screening process concentrates on just those technologies that are potentially applicable based upon the established criteria which includes remedial objectives, site specific conditions and the characteristics of the contaminants of concern. This section introduces and discusses the technologies in each general response action and presents the results of the screening process. Remedial technologies are discussed in the order in which they are listed in Table 6-3.

#### 6.5.1 No Action

The No Action response for the Sandoval Zinc site means that no remediation of impacted material, soil, groundwater or sediment will be designed or implemented. Under a No Action scenario, contaminants may leach from the soil and migrate to the groundwater. Contaminants may also migrate off-site through wind dispersion and surface water run off.

Although the No Action alternative does not remove or treat the sources of contamination, this general response action is required by the National Contingency Plan (NCP) and is retained to provide for a comparison of public health and environmental impacts later in the evaluation process.

### 6.5.2 <u>Institutional Controls</u>

Institutional Controls (ICs) represent minimal actions necessary to reduce the potential for exposure to the contaminants on-site. Two forms of ICs commonly used include: (1) Groundwater Monitoring and (2) Access restrictions.

Groundwater monitoring involves sampling and laboratory analysis of groundwater samples collected from existing monitoring wells. Monitoring can be implemented to determine whether the groundwater quality is deteriorating through contaminant migration.

Access restrictions are intended to reduce public access to the site and thus reduce the opportunity for exposure to contaminants. The IEPA has already implemented one form of access restrictions by installing a fence around the site and posting warning signs to restrict physical access to the site. Another form of restriction that could be imposed is deed restrictions. Deed restrictions may be used to restrict activities such as installation of drinking water wells, property resale and property use.

ICs can be considered as a part of most remedial alternatives, and are therefore retained for further evaluation.

#### 6.5.3 Containment

There are two containment technologies applicable to the Sandoval Zinc site: (1) Barrier Walls to contain movement of impacted groundwater and (2) capping to isolate impacted soils.

#### 6.5.3.1 Groundwater Containment Vertical Barriers

Impermeable barriers can be used to divert groundwater flow around the site or to contain impacted groundwater from migrating off-site. Various methods and materials considered for use in constructing groundwater barriers include the following:

- Slurry walls
- Grout curtains

### Sheet pilings

#### Slurry Walls

Slurry walls are the most common subsurface barriers utilized because they are a relatively effective method of reducing groundwater flow in unconsolidated soils. The slurry wall is constructed in a vertical trench that is excavated under a slurry. This slurry, which is usually a mixture of bentonite and water, acts essentially like a drilling fluid in that it hydraulically shores the trench wall to prevent high fluid losses into the surrounding soil. Slurry wall types are differentiated by the material used to backfill the slurry trench. Two of the most commonly used methods are: (1) soil-bentonite, and (2) cement-bentonite.

Soil-bentonite slurry walls are the most commonly used subsurface barriers. They can be installed either upgradient of the site to divert groundwater flow, downgradient to partially contain contaminant plumes or around the circumference of the site for containment. Soil-bentonite slurry walls are constructed by backfilling a vertical trench with a mixture of soil, bentonite and water. In the vertical perspective, the slurry wall may be either "keyed-in" or hanging. Keyed-in slurry walls are constructed in a trench which has been excavated into a low-permeability confining layer such as a clay deposit or bedrock. This layer will form the bottom of the contained site and a good key-in is essential to adequate containment. Hanging slurry walls, however, are not tied into a confining layer but extend several feet into the water table to act as a barrier to floating contaminants (such as oils and fuels) or migrating gases. The use of hanging slurry walls in site remediation is therefore, relatively rare and most installations utilize keyed-in slurry walls.

Soil-bentonite slurry walls have the lowest overall cost, the widest range of chemical compatibilities and the lowest permeabilities if properly constructed. At the same time, soil-bentonite walls have the highest compressibility (least strength), require a large work area, and because the slurry and backfill are fluid, they are only applicable to sites that can be graded to nearly level.

Cement-bentonite slurry walls share many of the same characteristics with soil-bentonite slurry walls. The principal exception is that the excavated trench is filled with a slurry composed primarily of portland cement and bentonite. Only a small percentage of the natural soils are also used in this mixture. The cement-bentonite slurry is allowed to set, forming a low permeability containment barrier. Generally less area is required for construction when compared to soil-bentonite slurry walls, however, excavated soils from the trench must be disposed of properly.

Slurry wall construction requires a large work area which may not be available at the Sandoval Zinc site. In addition, they are not effective unless keyed into a continuous confining unit. The lithology at the Sandoval Zinc site does not provide these conditions. Therefore, this technology is not retained for further evaluation.

#### **Grout Curtains**

Grout curtains are subsurface barriers that are constructed by injecting grouting material, under pressure, into the ground around the area to be contained. The grouting material can consist of cement, cement-bentonite slurry, alkali silicates, or organic polymers. The design of a grout curtain depends on soil characteristics and the capatability of the grout with the contaminant(s) to be contained.

Grout curtains are rarely applied to contaminated sites for many reasons. A major concern is that inadequate grout penetration could create gaps or discontinuities in the curtain. Grout curtains also require more monitoring than any other type of subsurface barrier and they may not be always capable of attaining very low permeabilities. Therefore, this technology is not retained for further consideration.

#### Sheet Piling

Sheet pilings are vertical metal or wood sheets driven into the ground to create a subsurface wall. They are usually installed to keep water out of a given construction area. The sheet piles are constructed by interlocking the sheet edges and driving them into the earth a short distance at a time until the desired depth is attained over the entire length of the wall. Sheet piling is used for temporary dewatering of an area, as well as for erosion protection, where the wall system would be subject to flowing surface water or wave action. The major parameters to be considered in the design of sheet piling are material permeability and the wall dimensions.

Two of the largest drawbacks of sheet piling are corrosion and the deflection of the piles by rocks or buried debris. This damage would likely render the wall ineffective and it is very difficult to inspect the completed structure for such damage. Therefore, due to the unpredictability of the integrity of the wall as well as the unfavorable lithology of the site, this technology is not retained for further evaluation as a groundwater barrier.

# 6.5.3.2 Capping

Capping technologies are used primarily to minimize the potential for direct contact with contaminants and reduce off-site transport of exposed contaminants and waste materials.

Caps containing impermeable barriers also minimize the percolation or infiltration of precipitation/surface waters. Capping can involve the installation of a compacted soil zone over the waste and can include an overlying layer of topsoil and vegetation cover. Excavation and/or regrading of some of the material in preparation for capping is also usually required.

The selection of capping materials and cap design is influenced by the remedial objectives as well as specific factors such as local availability and cost of cover materials, properties of cover materials, the nature of the contaminants being covered, local climate and hydrogeology, and the projected future use of the site in question. For the Sandoval Zinc site, three capping methods were considered: (1) a non-RCRA cap, (2) a RCRA cap, and (3) vegetation.

### Non-RCRA Cap

A non-RCRA cap contains just a single layer of low permeability material, and may be acceptable if there is reasonable assurance that the integrity of such a cap will be continually maintained. A drainage layer is usually not provided over the impermeable layer, so grading must be provided to convey water away from the cap. However, since the cap is made of material which is not impermeable surface water will still pass through. None the less, a non-RCRA cap will reduce, the risk of exposure through inhalation and ingestion of contaminants in the soil. Therefore, the non-RCRA cap is retained for further evaluation.

#### RCRA Cap

A RCRA cap generally contains two layers of impermeable materials to provide assurance of a long service life, and generally consists of an overlying drainage layer and an underlying foundation layer. The low permeability layer may consist of some combination of clay, cement, concrete, asphalt, or synthetic membranes. The drainage layer is designed to convey water away from the layer of low permeability thereby limiting the hydraulic head on the material and the potential for infiltration. Drainage and foundation layers are usually constructed of sand, crushed stone or geotextile drainage fabrics.

RCRA caps are normally used to cover highly contaminated areas in order to prevent infiltration and exposure to contaminants. This level of protection may be necessary to prevent the risk of exposure through inhalation, ingestion and direct contact. Therefore, this option is retained for further evaluation.

#### Vegetation

Vegetation is a special class of cap. Unlike a non-RCRA cap, no low permeability material (e.g., clay) is placed on top of the impacted soil. On the other hand, top soil is placed as a cover over impacted surface soil. A geotextile fabric may be installed to separate the clean topsoil cover and the impacted soil. The fabric may also provide an additional barrier through its resistance to excavation by small tools. Vegetation is induced by seeding the top soil with appropriate plant species. Deep rooted vegetation, which may threaten capping systems, should be avoided. Vegetation is aesthetically appealing and protects the soil cover from erosion. Vegetation is retained for further evaluation.

# 6.5.4 Pump-and-Treat

Groundwater pump-and-treat systems involve the extraction of impacted groundwater and treating the recovered groundwater above ground to remove the contaminants of concern. This technology involves the installation of extraction wells or collection trenches and submersible pumps to extract the groundwater for treatment.

The feasibility of treating impacted groundwater is dependent on the contaminants present, their concentrations, the physical/chemical properties of the contaminants in the groundwater, and the properties of water bearing unit.

Once the groundwater has been extracted there are several technologies available which can be utilized to treat the water. These treatment systems include both physical and chemical treatment. Physical treatment systems which were considered include filtration, reverse osmosis and ion exchange. The only chemical treatment technology evaluated was chemical precipitation.

#### 6.5.4.1 Physical Treatment

Physical treatment removes contaminants from the groundwater through processes that involve only a physical change. Dissolved metal salts are the contaminants of concern in the groundwater at the Sandoval site. However, the dissolved metals can be adsorbed by suspended solids. These dissolved and suspended solid contaminants can be separated from groundwater to a different medium. The commonly used technologies to affect this transfer are filtration, reverse osmosis and ion-exchange.

#### **Filtration**

Filtration is a process of separating and removing suspended solids from a liquid by passing the liquid through a porous medium. The porous medium may be fibrous fabric (paper or cloth), a screen or a bed of granular material such as sand. Suspended solids are not of primary concern at the Sandoval Zinc Site. However, the dissolved metal salts can become associated with the suspended solids and pretreatment by filtration is appropriate to prevent plugging or overloading of downstream process equipment used for the removal of the metal salts. Filtration is effective for removing suspended solids before treatment or removing flocculants after metals precipitation, and is retained for further evaluation.

#### **Reverse Osmosis**

Osmosis is when a semi-permeable membrane separates two solutions of different dissolved solids concentrations, pure water will flow through the membrane into the concentrated solution, while ions (e.g. dissolved salts) are retained behind the membrane. During reverse osmosis (RO), pressure is applied to the more concentrated solution to reverse the normal osmotic flow, and pure water is forced through the semi-permeable membrane into the less concentrated solution. The three most commonly used RO membrane materials are cellulose acetate, aromatic polyamides, and thin-film composites (consisting of a thin film of a salt-rejecting membrane on the surface of a porous support polymer). The membrane utilized for any particular system is dependent on temperature, pH and other limitations of the membrane material.

RO is primarily used to separate water from a feed stream containing inorganic ions. The purity of the recovered water is relatively high, and the water is generally suitable for recycling. The maximum achievable concentration of salt in the reject stream is usually 100,000 ml/L because of osmotic-pressure considerations.

One of the major applications of RO has been in the electroplating industry. The separation process does not require a energy intensive phase change and a result operating costs associated with energy consumption are relatively low. Capital costs are also relatively low and a low degree of operational skill is required. Therefore, this remedial technology has been retained for further evaluation.

#### Ion Exchange

Ion exchange is a separation process in which selected pollutant ions in a wastewater are removed by the ion exchange material (resin), while non-pollutant ions are exchanged from the resin into the wastewater. In practice, ion exchange "beads" are placed in a column and

water to be treated is passed through the bed. Most ion exchange resins are high-molecularweight organic polymers onto which chemical functional groups (e.g., sulfonic, carboxylic, phenolic, amines) are added.

The degree of the reaction (exchange) will depend on the resin's selectivity and as a separation technology, ion exchange does not eliminate the ionic contaminants but concentrates them. The saturated resin must be replaced or regenerated after each loading cycle.

Ion exchange has been used for the purification of public water supplies and demineralization (softening of water in process industries, particularly in metal plating and electronics manufacturing. Ion exchange systems are available and can be easily fabricated for specific applications and thus have been retained for further evaluation.

#### 6.5.4.2 Chemical Treatment

Chemical treatment involves removing contaminants from the groundwater through chemical change. The most commonly available technology applicable for chemical treatment of groundwater impacted with heavy metals like at the Sandoval Zinc site is chemical precipitation.

#### **Chemical Precipitation**

Precipitation is a process by which the chemical equilibrium of a waste stream is altered to reduce the solubility of heavy metals. The metals precipitate out as a solid phase and are taken out of the solution by solids removal processes. Metals precipitation is not one unit operation but a combination of coagulation, flocculation, sedimentation, and filtration processes.

The solubility of most heavy metals is reduced by raising the pH of a wastewater from 8 to 12. Although removal of metals as sulfides or carbonates is effective, hydroxide precipitation is, by far, the most common precipitation process. In hydroxide precipitation, hydrated lime (i.e., calcium hydroxide) or caustic (i.e., sodium hydroxide) is added for pH adjustment. Both alkalies have advantages and disadvantages. The cost of lime is less than that of caustic; however, the feed equipment is more expensive. Lime also produces a drier cake than caustic but sludge production is greater.

Adjustment of pH alone, however, is usually insufficient for removal of the insoluble metal hydroxide solids. Coagulants, such as iron salts, alum, and polymers, must be added to neutralize charges and to cause the formation of metal precipitates. Chemical coagulants

are added in a rapid mix tank and are followed by gentle mixing or "flocculation," which causes interparticle bridging and formation of flocs which settle rapidly. The settled solids can then be removed by a clarifier, a filter, or both.

Metal hydroxide precipitation is an established wastewater treatment process for the electroplating and metal finishing industries. Therefore, this technology was retained for further evaluation.

#### 6.5.5 Soil Treatment

Soil treatment technologies applicable to the Sandoval Zinc site are divided into two categories: (1) physical/chemical treatment technologies and (2) solidification/stabilization technologies.

### 6.5.5.1 Physical/Chemical Treatment

Physical treatment consists of transferring the contaminants in the soil to another media. Chemical treatment removes the contaminant through chemical reaction. A brief discussion of applicable technologies follows.

#### **Chemical Extraction**

This process involves mixing the impacted soil with a concentrated acid or chelating solution. The acid solution extracts the metals from the soil which is then thoroughly washed and returned to its original location. However, a large portion of the impacted soil at the site consists of slag from the smelting process. The slag contains high levels of heavy metals and is not easily reduced in size to expose the metals for extraction. Therefore, chemical extraction is not retained for further consideration.

#### Electro-Reclamation

Electro-reclamation removes heavy metals and other contaminants from soil and groundwater based on the phenomena of electro-osmosis, electrophoresis and electrolysis. These phenomena occur when the soil is electrically charged with direct current (DC) by means of one or several electrode arrays. Metal contaminants migrate to the negatively charged electrodes and are captured in the chemical solution circulating in the electrode. The solution is then treated in a water treatment facility.

Electro-reclamation can be applied both in-situ and on excavated soil. Bench scale experiments on fine sand (Geokinetics, 1989) have shown that cadmium concentrations can

be reduced from 319 mg/kg to less than 1 mg/kg (> 99% removal). Lead concentrations were reduced from 638 to 238 mg/kg (65% removal). Other soil types were also tested, but had lower removal efficiencies. Field experiments were conducted on a sediment layer (70 m long x 3 m wide x 20-50 cm deep) impacted with lead and copper. Lead removal efficiencies varied from 50-94 percent with an average of 74 percent. Other field experiments have also been conducted to evaluate removal of metals such as zinc, cadmium, and arsenic with varying degrees of success.

The subsurface soils at the Sandoval site consist of silt clays which do not have a high hydraulic conductivity. As a result, recover efficiencies are not expected to be high and this remedial technology was not retained for further evaluation.

#### Soil Washing

The soil washing process extracts contaminants from soil using water or an aqueous solution composed of chelating agents, surfactants, acids, or bases. The primary function of soil washing is a physical volumetric reduction of fine silt, clay, and colloidal fractions from cleanable coarse sand and gravel components, since the fine silts and clay typically absorb organic contaminants.

This technology has been demonstrated to remove halogenated and nonhalogenated hydrocarbons and heavy metals such as lead, cadmium, chromium copper, and nickel. This technology is most effective for soil with a high proportion of sand having a majority of soil particles greater than 200 mesh, or 0.074 mm (USEPA, 1988b). The subsurface soils at the Sandoval site consist primarily of slag. As a result the metals associated with the impacted soil have not been adsorbed but are inherent to the soil. Therefore, although soil washing will remove some of the metals in the soil it will not effectively remediate the soil. This technology was therefore not retained for further evaluation.

### 6.5.5.2 Solidification/Stabilization

Two types of solidification/stabilization technologies are applicable to the Sandoval Zinc site: (1) On-site stabilization and (2) In-situ stabilization.

#### On-Site Stabilization/Solidification

On-site stabilization methods are designed to immobilize contaminants, minimize leaching potential reduce toxicity of the waste, and improve the waste handling characteristics. Impacted material is excavated and mixed with treatment reagents that combine physically and/or chemically with impacted materials to decrease the mobility of the waste

constituents. The end product may be a standing monolithic solid or may have a crumbly, soil-like consistency, depending upon the amount and type of reagent added. After the contaminant is immobilized, the material can be consolidated to a common area of the site and placed in on-site containment or an engineered landfill.

On-site stabilization has demonstrated full-scale success as a remediation technology for the treatment of wastes such as the soils and sediment at the Sandoval Zinc site which contain heavy metals. This technology, however, will increase the volume of soil or sediment substantially and is therefore, only retained for further consideration to remediate the impacted sediments of the farm pond.

### In-Situ Stabilization/Solidification

As with on-site stabilization, in-situ stabilization methods are designed to immobilize contaminants, minimize leaching potential, and reduce toxicity of the waste. With in-situ stabilization, impacted soil is left in-place and mixed with treatment reagents to decrease the mobility of the waste constituents. Stabilization continues throughout the impacted area until all contaminants of concern are immobilized.

In-situ stabilization can effectively immobilize wastes containing heavy metals, PCBs, and PAHs with high molecular weight. The amount and type of reagent used is determined by the contaminants of concern, their respective concentrations, and the soil type. The use of in-situ stabilization would require several formulation of reagent. However, greater process control is afforded by excavating the material.

Soil mixing is divided into two categories, Shallow Soil Mixing (SSM) and Deep Soil Mixing (DSM). The SSM system uses a crane mounted mixing system. The mixing auger, three feet to 12 feet (1.0 meter to 3.7 meters) in diameter, is driven by a high torque turntable. The mixing head can be enclosed in a bottom-opened cylinder to allow for closed system mixing of the waste and treatment chemicals.

Treatment chemicals are transferred pneumatically for dry chemicals or pumped in cases where fluid chemicals would be used. Treatment chemicals are precisely weighed (for dry systems), or volumetrically measured (for fluid systems), to allow the correct proportions to be mixed with the untreated waste sludge or soil. The bottom-opened cylinder is lowered into the waste and the mixing blades are started while chemicals are introduced. The blades mix through the total depth of the waste in an up-and-down motion. A negative pressure is kept on the head space of the bottom-opened cylinder to pull any vapors or dust to the vapor treatment system. Once a cylinder of waste is mixed, the blades are retracted inside the bottom-opened cylinder and the cylinder is removed. The cylinder is then placed

adjacent, and overlapping, to the previous cylinder and the process is repeated until all waste has been treated.

In-situ stabilization is potentially applicable to the Sandoval Zinc site, and is retained for further consideration.

## 6.5.5.3 Metals Recovery

Metals recovery from soil appears to be feasible because high levels of zinc (10,000 - 100,000 mg/kg), lead (1,000 - 10,000 mg/kg) and copper (> 1,000 mg/kg) exist in the subsurface soils. Metals can be recovered by using thermal processes or by heap leaching.

#### Thermal Processes

Thermal processes involve concentrating the metals concentration in the soil by physical/chemical treatment-processes (e.g., air floatation, chemical extraction, chemical oxidation) and then recovering the metals in high temperature furnaces. Recovery of metals from smelter residues is a common practice in the mining industry. Therefore, thermal processes are retained for further evaluation.

#### Heap Leaching

Heap leaching is a commonly used technique in the mining industry to recover valuable metals from slag or tailings generated from primary processing of ores. The technique consists of constructing a heap of the material and leaching the heap with a suitable reagent. The heap is constructed on an impervious pad with a system for collecting the leachate, which is then recycled. The commonly used reagents which may be appropriate for removing zinc and copper from the tailings/slag are sulfuric acid, potassium cyanide, and nascent chlorine solution.

Heap leaching has the following advantages:

- The technique is demonstrated and proven effective for recovering valuable metals like gold and silver.
- The system is simple to construct and install.
- The cost of processing is typically low compared to other above ground thermal recovery processes.

Although the characteristics of the soil are likely to be different from slag or tailings, heap leaching is potentially applicable and retained for further evaluation.

# 6.5.6 Excavation and Removal

Removal technologies refer to methods used to excavate and handle soils, sediments, wastes and solid materials. Excavation technologies provide no treatment of the wastes, but may be used prior to treatment or disposal technologies to facilitate removal of wastes from designated areas. Dewatering and supernant treatment may also be conducted in conjunction with removal technologies.

#### 6.5.6.1 Soil Removal

Excavation of contaminated soils or subsurface wastes may be performed by a variety of technologies. Typical equipment includes draglines, loaders, dozers, pans (scrapers), backhoes and trucks. Excavated material can be loaded onto trucks, and hauled off-site to an approved treatment and disposal facility or it can be treated and disposed of on-site.

This technology was therefore retained for removal of soil and sediment.

# 6.5.7 Disposal

Disposal options available for the Sandoval Zinc site vary according to the media. Options for each of the media present at the site are discussed in the following paragraphs.

# 6.5.7.1 Groundwater Discharge

Three process options are considered for groundwater discharge: (1) off-site discharge to a Publicly Owned Treatment Work (POTW) for treatment, (2) recharge after treatment and (3) discharge after treatment to surface water body.

# Off-Site Discharge

Off-site disposal involves extracting the impacted groundwater and transporting it off-site to a POTW for treatment. The City of Centralia, which supplies water for Sandoval, has a wastewater treatment plant. This plant is located close to the Sandoval Zinc site. However, since the sludge generated by the POTW is used for land applications, the POTW is not permitted to accept groundwater impacted with metals. Therefore, this option is not retained for further evaluation.

# Recharge to Groundwater

Subsurface distribution systems, such as french drains, infiltration galleries or injection wells, are means of returning the treated groundwater at shallow depths. If the water is discharged within the site boundaries it could flush out more contaminants, thereby increasing treatment requirements, or dilute the existing impacted water, thereby reducing the efficiency of the treatment system. Therefore, this option is eliminated from further consideration.

### Discharge to Surface Water Body

If the treated groundwater complies with Illinois State water quality standards, it could be discharged to the off-site drainage ditch. Therefore, this discharge option is retained for further evaluation.

### 6.5.7.2 Sludge Disposal

Groundwater treatment will generate sludge containing the metals removed from impacted groundwater. The quantity of sludge generated is expected to be small and can be discharged off-site to a permitted RCRA facility. This option is retained for further consideration.

# 6.5.7.3 Excavation and Land Disposal

Excavation and land disposal is an established and commonly used technology for impacted soils. Two types of available disposal options are: (1) off-site secure landfill and (2) on-site secure landfill.

#### Off-Site Secure Landfill

In this option, impacted soil would be excavated and transported to an off-site RCRA disposal facility. This option is retained for further evaluation.

#### On-Site Secure Landfill

In this option, a secure landfill is constructed on-site for the disposal of impacted soil. This option is impractical because the site does not have sufficient area to construct a secure landfill. In addition, the shallow depths to groundwater (average 5 ft.) on-site further limits the area available for disposal. Therefore, on-site landfill disposal is not retained for further consideration.

#### 6.5.8 Collection

Collection technologies are an integral part of any groundwater treatment system. Numerous structures or mechanical systems can be used to collect and transfer impacted groundwater for treatment.

# 6.5.8.1 Subsurface Drainage Systems

Although many different types of subsurface drainage systems are commonly used, only one system is described due to its applicability based on the limited area and site specific conditions.

### French Drain/Interceptor

French drains and interceptor trenches are two subsurface drainage systems that can be used to collect or intercept and convey groundwater by gravity flow. They can serve the same general purpose as a groundwater pumping system as they create a continuous zone of influence in which groundwater flows toward the drains.

The drains are typically placed perpendicular to the direction of the groundwater flow to intercept the contaminant plume or prevent groundwater movement into a impacted area. The drains are constructed by excavating a trench, lining the trench with filter fabric, placing a gravel bed with perforated drain pipe, and backfilling the trench with gravel. Intercepted groundwater flows along the trench or french drain to a collection point or sump for discharge and/or treatment and discharge.

Construction of a french drain or interceptor trench would involve the excavation and disposal of potentially impacted soils. The presence of a shallow water table and fine sands would require that excavation be sheeted and braced to limit excavation quantities and assure stability of existing structures. This process option is retained for further evaluation.

#### 6.5.8.2 Diversion

Diversion involves regrading the site to reduce surface water infiltration and control erosion. Grading is often performed as part of surface scaling activities. Grading is therefore retained for further evaluation.

# 6.6 Disposal of Waste Product/Ash and Debris

Off-site disposal to a secure RCRA landfill is a viable technology for disposing the waste product/ash and debris stored on-site. This technology is retained for further evaluation.

# 6.7 Disposal of Above Ground Storage Tank

The above ground storage tank which contained fuel oil requires proper closure. This action consists of decontaminating the tank and associated piping, properly disposing of rinse waters generated during the decontamination, and appropriately disposing of the tank offsite. Two options are available for disposal: (1) the tank can be removed and disposed offsite and (2) the tank can be abandoned in place. Both are potentially applicable to the site. Disposal is retained for further evaluation.

# 6.8 Summary of Initial Screening

Figure 6-2 summarizes the initial screening of technologies applicable to the Sandoval Zinc site. Technologies which are eliminated from further consideration are shaded.

# SITE SKETCH

STATE OF ILLINOIS

Date of Inspection: August 29, 1991

Inspector: Bruce Ford

**ENVIRONMENTAL PROTECTION AGENCY** 

Site Code: L1210500002

County: Marion

Site Name: Sandoval Zinc

Time: 1:00 pm - 2:25 pm

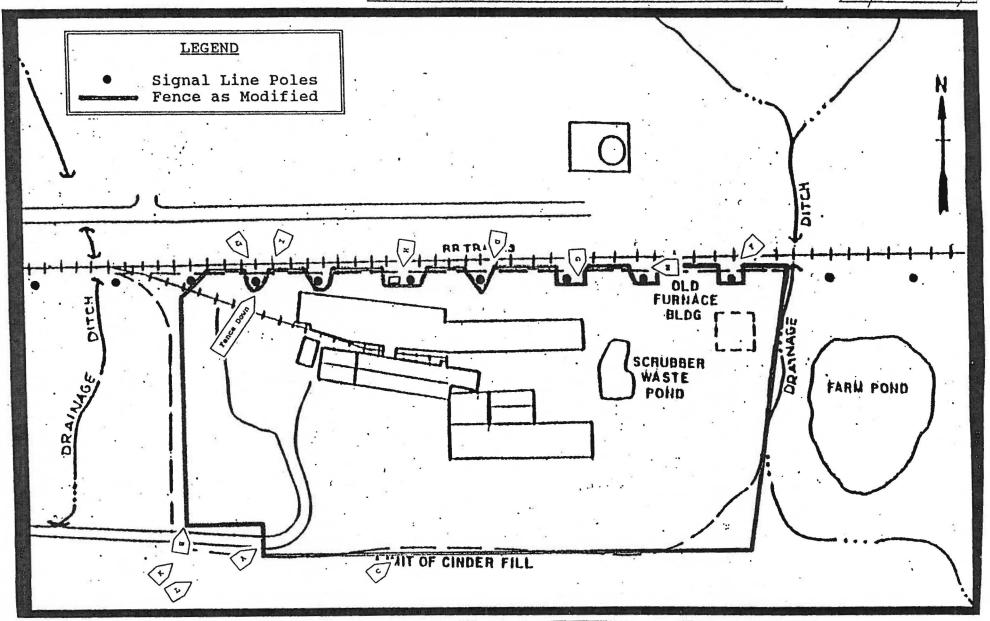
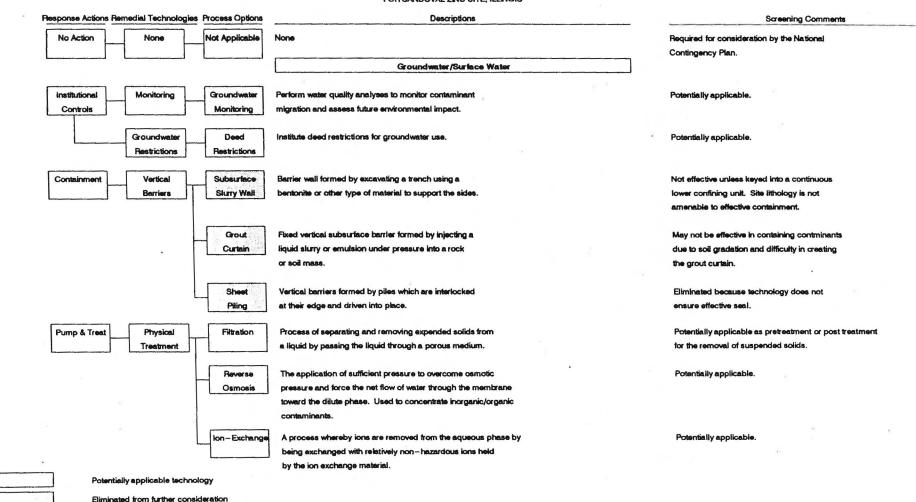


Figure 6-1: Study Area For The Feasibility Study

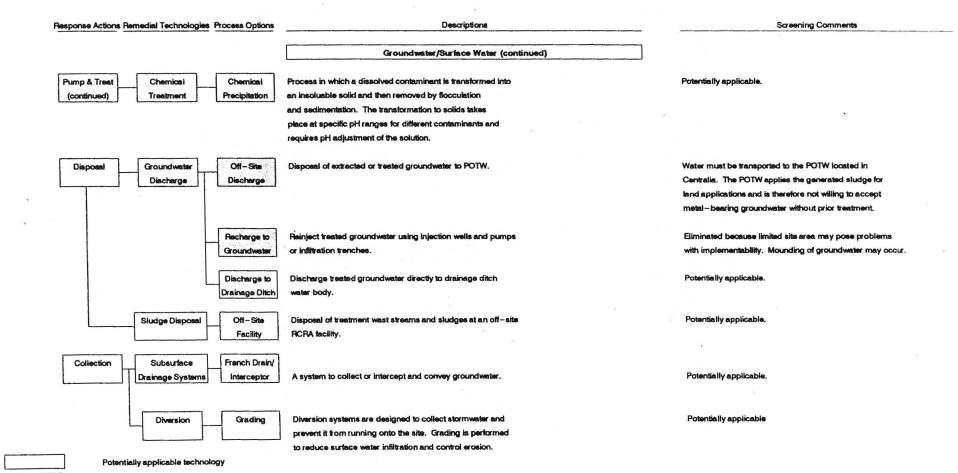
# FIGURE 6-2 INITIAL SCREENING OF TECHNOLOGIES AND PROCESS OPTIONS FOR SANDOVAL ZINC SITE, ILLINOIS

Page 1 of 6



Eliminated from further consideration

# FIGURE 6-2 INITIAL SCREENING OF TECHNOLOGIES AND PROCESS OPTIONS FOR SANDOVAL ZINC SITE, ILLINOIS



#### FIGURE 6-2 INITIAL SCREENING OF TECHNOLOGIES AND PROCESS OPTIONS FOR SANDOVAL ZINC SITE, ILLINOIS

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nse Actions Remedial Technologies Pr	ocess Options	Descriptions	Screening Comments
		Soil	
——————————————————————————————————————	Groundwater Monitoring	Perform water quality analyses to monitor contaminant migration and assess tuture environmental Impact.	Potentially applicable.
Access Restrictions	Deed Restrictions	Institute deed restrictions for impacted property.	Potentially applicable
	Fencing	Fence impacted property to isolate the site and minimize direct contact with contaminated soils.	This has already been inplemented by IEPA.
ntainment Capping	RCRA Cap	A cap that conforms to RCRA design criteria covering the soil to eliminate infiltration of precipitation and eliminate risk of exposure through inhalation and ingestion.	Potentially applicable
-{	Non-RCRA Cap	A cap of low permeability material to minimize infiltration of precipitation and reduce risk of exposure through inhalation and ingestion.	Potentially applicable.
	Vegetation	Consists of covering the surface soil on site with soil and seeding the soil for vegetation to minimize erosion and reduce risk of exposure through inhalation or ingestion.	Potentially applicable.

Potentially applicable technology

Eliminated from further consideration

Eliminated from further consideration

# FIGURE 6-2 INITIAL SCREENING OF TECHNOLOGIES AND PROCESS OPTIONS FOR SANDOVAL ZINC SITE, ILLINOIS

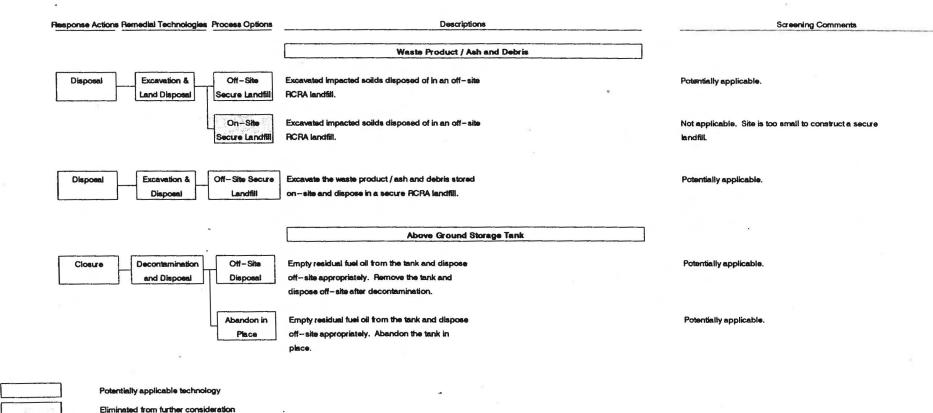
Page 4 of 6

Response Actions Remedial Technologies Process Options Descriptions Screening Comments Soil (continued) Treatment of soil with acid solutions to remove metals. Treatment Physical/Chemical Chemical Not feasible because the majority of impacted soil consists Treatment Extraction of slag which contains high levels of heavy metals. The slag is also not easily reduced in size to expose the metals for extraction. Electro An electric current is passed through electrodes imbedded The subsurface soils at the site consist of silty clays which do not Reclamation in the soil. Metal contaminants migrate to negatively have a high hydraulic conductivity. As a result, recover efficiencies charged electrodes and are captured in the chemical solution are not expected to be high. circulating in the electrode. Removes contaminants from soils using a washing fluid with Subsurface soils at the site consist primarily of slag. As a result Soil Washing appropriate surfactants, acids or chelating agents. the metals associated with the impacted soil have not been adsorbed but are inherent. Therefore, soil washijng would not be effective.. Solidification/ On-Site Excavated impacted soils are mixed with stabilizing Impractical because the volume of resultant product will Stabilization Stabilization/ agents and other additives above ground to produce a increase substantially over the original volume. Solidification stabilized material. The process immobilizes contaminants within a solid matrix. In-Situ Stabilizing agents are put directly into the impacted Potentially applicable. Stabilization/ soil through a rotating shaft. At the end of the treatment, Solidification a treated block of soil remains. The process immobilizes contaminants within a solid metrix. Soils excavated and thermally treated to recover metal. Thermal Potentially applicable. Metals Recovery Processes Heap Soils excavated and leached with suitable reagents Potentially applicable. Leaching in heaps on impervious pads. Potentially applicable technology

# IONS

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# FIGURE 6-2 INITIAL SCREENING OF TECHNOLOGIES AND PROCESS OPTIONS FOR SANDOVAL ZINC SITE, ILLINOIS



#### FIGURE 6-2 INITIAL SCREENING OF TECHNOLOGIES AND PROCESS OPTIONS FOR SANDOVAL ZINC SITE, ILLINOIS

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Response Actions Remedial Technologies Process Options	Descriptions	Screening Comments
	Farm Pond Sediments	
		*
Treatment Solidification/ On-Site Stabilization Stabilization/	Excavated impacted sediments are mixed with stabilizing agents and other additives above ground to produce a	Potentially applicable.
Solidification	stabilized material. The process immobilizes contaminants within a solid matrix.	
In-Situ Stabilization/ Solidification	Stabilizing agents are put directly into the impacted sediment using backhoes. At the end of the treatment, a treated block of sediments remains. The process immobilizes	Potentially applicable.
	contaminants within a solid matrix.	
Disposal Excavation & Off-Site Secure Disposal Landfill	Excavate the impacted sediments and dispose in a secure RCRA landfill	Potentially applicable.

Potentially applicable technology

Eliminated from further consideration

Eliminated from further consideration

# FIGURE 6-3 EVALUATION OF PROCESS OPTIONS Sandoval Zinc Site, Illinois

Page 1 of 6

Response Actions Remedial Technologies **Process Options** Effectiveness Implementability \* Cost No Action None Not Applicable Does not achieve Remedial Action Objectives (RAO's). Not applicable. None GROUNDWATER Institutional Monitoring Groundwater Does not achieve Remedial Action Objectives (RAO's) Readily implementable. Low Controls Monitoring May be used in conjunction with other process options. Groundwater Deed Does not achieve RAO's. May be used in conjunction with May be acceptable to local public or government Nominal Restrictions Restrictions other process options. agencies with additional process options. Pump & Treat Filtration Effective for removal of suspended solids. Should be coupled with **Physical** Readily implementable. Low another technology for pre or post treatment process. Treatment Reverse Effective for removal of metals. Process creates brine waste Readily implementable. Large volume of brine High stream that may require treatment. Osmosis waste requiring treatment is generated. ion-Exchange Effective for removal of metals. Resins are selective and may Readily implementable. Concentrated spent High remove multiple ions. Process creates regenerate solutions regenerate is created which requires treatment. that require treatment. Chemical Chemical Effective and reliable for metals removal. Process creates sludge Readily implementable. Low to Precipitation that requires sludge disposal. Moderate Treatment Retained as a representative process option \* Groundwater \* Soil Low Cost = 0.20 - \$1.00/1000 gallon Low cost = \$20-\$70/cubic yard

Moderate Cost = \$70-\$120/cubic yard

High Cost = >\$120/cubic yard

Moderate Cost = \$1-\$5/1000 gallon

High Cost = \$5/1000 gallon

sand2.wk1

# FIGURE 6-3 (Continued) EVALUATION OF PROCESS OPTIONS Sandoval Zinc Site, Illinois

Page 2 of 6

Response Actions Rem	nedial Technologies Process Options	Effectiveness	Implementability	* Cost
	*	GROUNDWATER (continued)		
Disposal	Groundwater Discharge to Drainage Ditch	Effective in discharge of treated groundwater.	Implementable.	Low
	Sludge Disposal Off-Site RCRA Facility	Effective in disposal of treatment wastestreams and sludges.	Implementable.	High
Collection	Subsurface French Drain/ Interceptor	Effective in intercepting contaminated groundwater for treatment.	Implementable.	Low
	Diversion Grading	Effective in reducing surface water infiltration and controlling erosion.	Implementable.	Low

	metaned as a representative process op	
12 12	Eliminated from further consideration	

\* Groundwater

Low Cost = 0.20 - \$1.00/1000 gallon

Moderate Cost = \$1 - \$5/1000 gallon

High Cost = \$5/1000 gallon

\* Soil

Low cost = \$20-\$70/cubic yard

Moderate Cost = \$70-\$120/cubic yard

High Cost = >\$120/cubic yard

# FIGURE 6-3 (Continued) EVALUATION OF PROCESS OPTIONS Sandoval Zinc Site, Illinois

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Response Actions Remedial Technologies Process Options Effectiveness Implementability \* Cost SOIL Institutional Monitoring Groundwater Does not achieve RAO's. May be used in conjunction Readity implementable. Low Controls Monitoring with other process options. Deed Does not achieve RAO's. May be used in conjunction Nominal Access May be acceptable to local, public or Restrictions Restrictions with other process options. government. RCRA Cap Effective in eliminating risk of exposure to contaminants Moderate Containment Capping Implementable. through inhalation, ingestion and direct contact. Non-RCRA Effective in reducing risk of exposure to contaminants Implementable. Low through inhalation, ingestion and direct contact. Сар Maintaining integrity of cap is a concern. Vegetation Effective in preventing soil erosion. Can be used in implementable. Low conjunction with capping to eliminate exposure to contaminants through inhalation, ingestion and direct contact

Retained as a representative	process option
	) <b>-</b>
Eliminated from further cone	ideration

\* Groundwater
Low Cost = 0.20 - \$1.00/1000 gallon
Moderate Cost = \$1 - \$5/1000 gallon

Moderate Cost = \$1 - \$5/1000 gallon High Cost = \$5/1000 gallon \* Soil

Low cost = \$20-\$70/cubic yard

Moderate Cost = \$70-\$120/cubic yard

High Cost = >\$120/cubic yard

sand2.wk1

# FIGURE 6-3 (Continued) EVALUATION OF PROCESS OPTIONS Sandoval Zinc Site, Illinois

Page 4 of 6

Response Actions Remedial Technologies Process Options Effectiveness Implementability \* Cost SOIL (continued) Physcial/Chemical Available performance data indicates that RAO's Treatment Chemical Has not been demonstrated on a commercial High Treatment Extraction are not likely to be achieved. scale. Electro Cadmium & lead concentrations cannot be reduced The technology is under development and is not High to established cleanup standards. Reclamation commercially available. Sol Based on available performance data, RAO's are not Lack of process control due to variations Low to Washing likely to be achieved. Extensive treatability studies in soil composition is a concern. Moderate required to determine optimum process conditions. Effective for long-term immobilization of the Solidification/ In-Situ Implementable Low to site in organic comtaminants. Stabilization Stabilization/ Moderate Solidification Thermal Processes effective for one metal may not be effective Metals Process efficiencies have to be tested to High Recovery Processes for another metal to the same degree. RAOs are not ensure performance. Large volumes of soil likely to be achieved for all metals of concern. require excavation and transportation off-site. Reagent effective for one metal may not be effective The area available on-site is too small to High Heap Leaching for another. Effective only during summer periods. implement this alternative. RAOs not likely to be achieved. \* Groundwater \* Soil Retained as a representative process option Low Cost = 0.20 - \$1.00/1000 gallon Low cost = \$20-\$70/cubic yard Eliminated from further consideration Moderate Cost = \$1-\$5/1000 gallon Moderate Cost = \$70-\$120/cubic yard High Cost = \$5/1000 gallon High Cost = >\$120/cubic yard

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sand2.wk1

### FIGURE 6-3 (Continued) EVALUATION OF PROCESS OPTIONS Condend The Site Illinois

Page 5 of 6

Sandoval Zinc Site, Illinois \* Cost Response Actions Remedial Technologies Process Options Effectiveness Implementability SOIL (continued) Off-Site Effective, but requires long-term management. There are no secure landfills close by. Disposal Excavation & Land Disposal Secure Landfill Off-site disposal may also be impractical because High of large volume of soil to be removed in order to achieve the RAO's. Waste Ash / Product & Debris Excavation Off-Site Secure Effective in eliminating risk of exposure through Implementable, aithough the waste must be Moderate Disposal transported a considerable distance. to High Landfill source removal. **Above Ground Storage Tank** Effective in eliminating risk of exposure Implementable Closure Decontamination Off-Site Low and Disposal Disposal through source removal. Effective in eliminating risk of exposure implementable Abandonment Low in Place assuming residual oil is properly disposed. Retained as a representative process option \* Groundwater \* Soil Low Cost = 0.20 - \$1.00/1000 gallon Low cost = \$20-\$70/cubic yard Eliminated from further consideration Moderate Cost = \$1-\$5/1000 gallon Moderate Cost = \$70-\$120/cubic yard High Cost = \$5/1000 gallon High Cost = >\$120/cubic yard

sand2.wk1

## FIGURE 6-3 (Continued) EVALUATION OF PROCESS OPTIONS Sandoval Zinc Site, Illinois

Page 6 of 6

Response Actions Remedial Technologies Process Options	Effectiveness	Implementability	* Cost
	FARM POND SEDIMENTS		
Treatment Solidification/ In-Situ Stabilization Stabilization/ Solidification	Effective for long-term immobilization of the site in organic comtaminants.	Implementable	Low to Moderate
On—Site Stabilization/ Solidification	Effective for long – term immobilization of the site in organic comtaminants.	Implementable	Low to Moderate
Disposal Excavation Off-Site Secure	Effective in eliminating risk of exposure through source removal.	Implementable, although the waste must be transported a considerable distance.	Moderate to High

 Retained as a representative process option
Eliminated from further consideration

\* Groundwater

Low Cost = 0.20 - \$1.00/1000 gallon Moderate Cost = \$1 - \$5/1000 gallon High Cost = \$5/1000 gallon

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\* Soil

Low cost = \$20-\$70/cubic yard

Moderate Cost = \$70-\$120/cubic yard

High Cost = >\$120/cubic yard

TABLE 6-1
SANDOVAL ZINC GROUNDWATER OBJECTIVES

	T T T T T T T T T T T T T T T T T T T	DWATER OBJECTIVES	T					
PARAMETER (mg/kg)	OBJECTIVE	BASIS	ADL					
Arsenic	0.2	wec <sup>1</sup>	0.01					
Cadmium	0.05	wac	0.002					
Chromium	1.0	Mac	0.01					
Copper	1.3	MGC	0.025					
Lead	0.1	MGC	0.005					
Mercury	0.01	wac	0.0002					
Nickel	2.0	wac	0.04					
Selenium	0.02	wac	0.005					
Zinc	10.0	WQC	0.02					
Manganese	10.0	MQC	0.015					
Barium	2.0	MGC	0.2					
Cobalt	1.0	Mac	0.05					
Benzene	0.025	MCL <sup>2</sup> & treatability	0.002					
Ethylbenzene	1.0	MCL & treatability	0.002					
Toluene	2.5	MCL & treatability	0.002					
Xylene	10.0	MCL	0.005					
Acetone	0.7	RfD <sup>3</sup>	0.01					
2-Butanone	0.35	RfD	0.1					
Naphthalene	0.039	RfD & treatability	0.01					
Acenaphthalene	2.1	RfD & treatability	0.018					
Anthracene	10.5	RfD & treatability	0.0066					
Flouranthene	1.4	RfD & treatability	0.0021					
Fluorene	1.4	RfD & treatability	0.0021					
Pyrene	1.05	RfD & treatability	0.0027					
Total Carcinogenic PNAs -benzo (a) anthracene -benzo (b) fluoranthene -benzo (k) fluoranthene -Chrysene -dibenzo (a,h) anthracene -indeno (1,2,3-c,d) pyrene	0.001	PMCL <sup>4</sup> & treatability	0.00013 0.00023 0.00018 0.00017 0.0015 0.0003 0.00043					
Total Non-Carcinogenic PNAs -acenaphthylene -benzo (g,h,i) perylene -phenanthrene	1.05	RfD for Pyrene	0.01 0.00076 0.0064					

WQC is the water quality criteria, USEPA 1972. Insufficient data are currently available for antimony, beryllium, silver and dibenzofuran. If contamination is detected following cleanup, COT should be contacted.

MCL is the maximum contaminant level, USEPA.
RfD is a reference dose, calculated by OCS.

PMCL is the proposed MCL

<sup>(1)</sup> Soil objectives for all heavy metals shall be based on an analysis using TCLP with results in mg/1.

TABLE 6-2: ESTIMATED MINIMUM VOLUMES OF SOIL FOR REMEDIATION

METALS IN SOIL	ESTIMATED MINIMUM <sup>2</sup> VOLUME OF SOIL TO BE REMEDIATED CUBIC YARD	ESTIMATED SOIL <sup>1</sup> CONCENTRATION AFTER REMEDIATION mg/kg	IEPA CLEANUP OBJECTIVES	BACKGROUND <sup>1</sup> SOIL CONCENTRATION
Cadmium	>330,555	, <1	0.05	0.04-1.5
Copper	> 94,444	<100	1.30	10-30
Lead	>141,667	<100	0.10	10-40
Zinc	>425,000	<100	10.00	20-50

Source: J.P. Gibb and K. Cartwright, "Retention of Zinc, Cadmium, Copper and Lead by Geologic Materials", Cooperative Groundwater Report 9, Illinois State Water Survey (ISWS) and Illinois State Geological Survey (ISGS), 1982.

Represents the minimum volume of soil requiring remediation to achieve the concentrations listed in column 3.

# Table 6-3 Feasible General Response Actions And Associated Remedial Technologies Sandoval Zinc Sandoval, Illinois

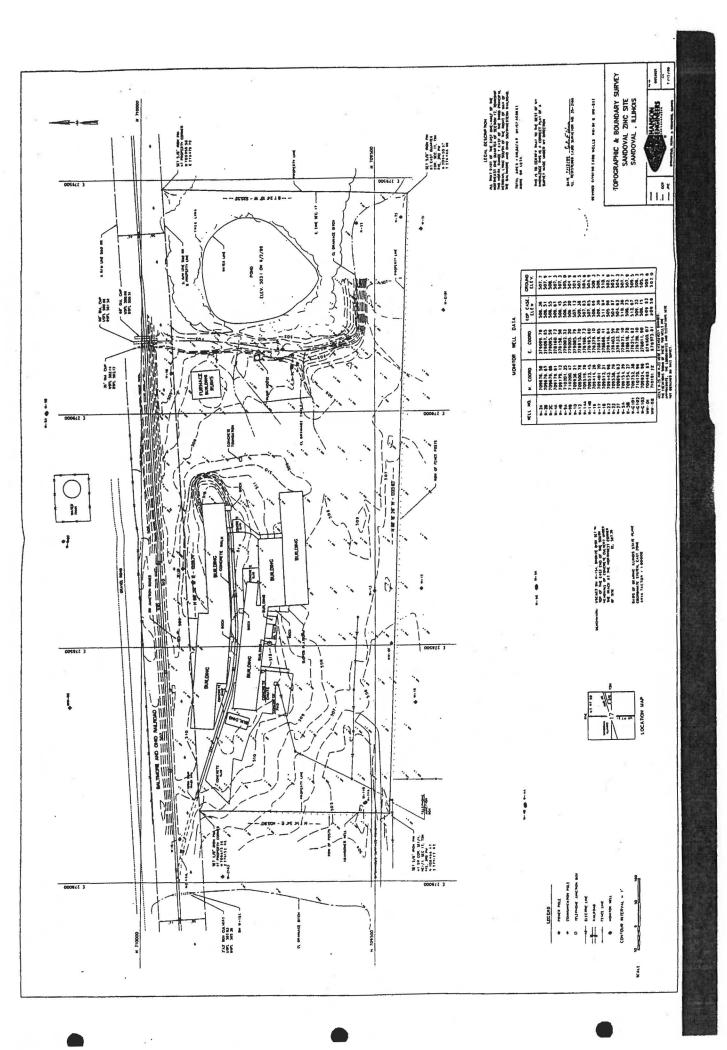
RESPONSE ACTION	REMEDIAL TECHNOLOGIES
No Action	- None
Institutional Controls	- Groundwater Monitoring - Access Restrictions
Containment	- Groundwater Containment Vertical Barriers - Capping
Pump & Treat	- Physical Treatment - Chemical Treatment
Soil Treatment	<ul><li>Physical/Chemical Treatment</li><li>Solidification/Stabilization</li><li>Metal Recovery</li></ul>
Excavation & Removal	- Soil Removal
Disposal	<ul><li>Groundwater Discharge</li><li>Sludge Disposal</li><li>Excavation and Land Disposal</li></ul>
Collection	<ul><li>Extraction Wells</li><li>Subsurface Drainage Systems</li><li>Diversion</li></ul>
Waste Product/Ash & Debris Disposal	- Off-Site Disposal
Above Ground Storage Tank Closure	- Off-Site Disposal

# Table 6-4 List Of Process Options Retained Sandoval Zinc Sandoval, Illinois

REMEDIAL TECHNOLOGY	PROCESS OPTION						
No Action	None						
GROUNDWATER/	SURFACE WATER						
Institutional Controls	- Monitoring - Deed Restrictions						
Pump & Treat	<ul><li>Filtration</li><li>Reverse Osmosis</li><li>Ion Exchange</li><li>Chemical Precipitation</li></ul>						
Collection	<ul> <li>Extraction Wells</li> <li>French Drain/Inteceptor</li> <li>Grading for diversion of Surface Water</li> </ul>						
Disposal	- Discharge to Off-Site Drainage Ditch						
	SOIL						
Institutional Controls	- Groundwater Monitoring - Deed Restrictions						
Containment	- Non-RCRA Cap - RCRA Cap - Vegetation						
Treatment	- Stabilization						
WASTE ASH/	PRODUCT & DEBRIS						
Disposal	- Off-Site Disposal						
ABOVE GROUND	STORAGE TANK						
Disposal	- Off-Site Disposal						
FARM POND	SEDIMENTS						
Institutional Controls	- Groundwater Monitoring - Deed Restrictions						
Treatment	- Stabilization						

#### 9.0 REFERENCES

- 1. J.P. Gibbs & K. Cartwright; <u>Retention of Zinc, Cadmium, Copper and Lead by Geologic Materials</u>, Cooperative Groundwater Report 9, Illinois State Water Survey, and Illinois State Geological Survey, 1982.
- 2. <u>Illinois Water Pollution Act</u>, Section 304.124, Subtitle C.
- 3. Ebasco Services Incorporated; <u>Information Summary and Data Assessment Report</u>, Sandoval Zinc Site, Sandoval, IL, April 14, 1989.
- 4. Hanson Engineers; Final Report of Survey Activities, Sandoval, IL, July 1989.
- 5. H.B. Willman, E. Atherton, T.C. Buschbach, C. Collinson, J.C. Frye, M.E. Hopkins, J.A. Lineback, & J.A. Simon; <u>Handbook of Illinois Stratigraphy</u>, Illinois State Geological Survey Bulletin 95, 1975.
- 6. H.B. Willman & J.C. Frye; <u>Pleistocene Stratigraphy of Illinois</u>; Illinois State Geologic Survey Bulletin 94, 1970.
- 7. M.J. Hvorslev; <u>Time Lag & Soil Permeability in Groundwater Observations</u>, U.S. Army Corps of Engineers Waterways Exp. Sta. Bulletin 36, Vicksburg, Mississippi, 1951.
- 8. R.A. Freeze & J.A. Cherry; Groundwater, Prentice-Hall, Inc., New Jersey, 1979.
- 9. U.S. Environmental Protection Agency; <u>Guidance for Conducting Remedial Investigations and Feasibility Studies Under CERCLA</u>, Interim Final, EPA/540/G-89/004, October 1988.
- 10. R.S. Means Company Inc., Means Heavy Construction Cost Data, 5th Annual Edition, 1991.



#### APPENDIX B

MONITORING WELL BORING LOGS &
WELL CONSTRUCTION DIAGRAMS

BORING LOG SHEET \_ PROJECT NO. 8652,102 PROJECT NAME Sandoval Z: DATE 6/14/90 PROJECT LOCATION Sandous MWOI WATER ENTERS DRILLED BY TEPA LOGGED BY K. WEBB ELEVATION 505.8 SPECIAL NOTES AND U.S.C. FIELD OBSERVATIONS SAMPLE DESCRIPTION REC RESIST TYPE oppn 0-85 Topsoil - med brown, sandy, 2.7 moist. 2.7 damp, light brownish gray med. SOA to 1.5, Abd. roots, mothed tappearance. 2.7-27/ 1.5-2.5 Silty clay, mottled gray and brown, 26d. 16/2ck carbonaceous Oppm OVA 5.5 material, firm, damp, less black 5.5 oppm material, friable. 4.2 3.6-55grades into sandy clay, high plasticity, met brownish 10 gray, less mottling damp, -5.5-10,05andy clay, same as above, scarce pebbles, becoming wet 21 base. 8.51 color becomes reddish brown from oxidation, 10. oppm OVA. med prasticity, becomes more 10 gravely with acpth. Oxidation gives mottled appearance to sample 10.0 - 11.95 Clayer sand, reddish brown to gray (mothed appearance wet, low plasticity, abd. gravel. bottom A' is light gray (no mottling).
-11.95.13.75 stayey sand, reddish brown, 125' is band of gravel, met, compact sand, 26d, gravel throughout, no mothling, mod. firm brown to med gray from, abd gravel, friable, Gottom of sample appears to be cleaner sand OPPM 15.0-18,3 - clayer sand - sandy clay, medigray, hard, friable (some as 13.75-15:0) Add gravel, 165-16.8 is sand pocket, low plasticity

poorly sorted,

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			v •,		BORING LOG	<b>.</b> SHI	EETOF
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		E ELI	EVATION		7.0 / ELEVATION DATUM	 r	SPECIAL NOTES AND
	DEPTH		SAMPLE		DESCRIPTION	U. S. C.	FIELD OBSERVATIONS .
	0-	TYPE	REC	RESIST			OVA=3.1 TOP
		G.	cil		0-1.1 dak brown top soil, loany		
		Confinuous	5/5		0.9-1.2 sandy layer, 1.2-5 gray ofty		-
		u O U	( )		- clay, high plasticity, moist, orango mettling coxidations) car bonecens metil	,	_
	2	Ċ-			disseminated throughout. Abundant roofs  0-705 decreasing withing with depth _  No HCL reacher 0-5		
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	10'-	1	4.8		10-12.8 orange - brown silty clas		OUA = 0.4
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			7		-and -mel		Free water -
				d, s	12.8-13.5 silty sand with some clay,		<b>-</b>
		1 :	चं~्त-		- gray with some orange staining.		
			7,		13.5-14,4: sity clay, gray with coursemble	*	OUA = 0-3
					orange starning high plasticity of and clay. 14.4. 15 arange sand sine above.		0VA=0.8
	13		, ,		15-18 - hown true gravel and course sant,		, –
		İ	24		non-cokes ve very wet self-rounded to -		Free weter
			13		founded graves. He leageneous lifel-gies,		0,2-01A
	1	_			for the clasts con for tout on y down and		CUA=0
	10	-			19-70 gray silty cay with some		X -
	-	V	el.		- the gravel. High planticity, mist		OUA-O
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					with overlying sand. Strong Heli reaction -		1
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					- After completion driller said the		
					bottom of the hole was probably more -	.	. 4

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COMPLETION DEPTH   SEDROCK DEPTH   TOP OF CASING   1:50 PM SCREEN	2"PVC, 5, 018LDT
11124 CASING	PERSONNEL
PVC, SCH 40, 5'SCREEN, 10, 5'	L. KEVIN ROSERS
SCREEN WITERVAL	O. KEN BOSIE
+2-17 11.7-16-5 PUC .015LDT	THE DALE HALFORD
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70-85 SUTY CLAY GRAY-RED BR-BL.	
TRAKE SAWA, LEHCKEN, OX	
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SANDY FEW PEBBLES TO	
i	
= 11.2-11.5 CLAY, GRAY, SANDY [1]	
F"¬	
3-11:5-12:5 SANA, POBBLEY, FINE MED GRAIN []	
TAN-BR SAT GRAY	
= 12.5-12.8 CLAY, ERAY, SANDY, CALC.	1
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12.8-16.7 TILL, GRAY BR. BLACK SAND, PLBBLES	



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7,	12.8-12.6 CLAY	, BR. DARK, V.TIN	E PHTRIX PLEBUES		7	-					
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1 2	m ARC	04 /	210500002		1				'  ^	SAL SADS	BENTONITE PELLETTS
1 "	SANDO	WAL/ZINC			4/1/8	7	4	3/8	7		10.35-11' BUCK
BORENG LO	CATION				1 ' '		•	•	- 1		
	NORTH O	F PLANT NEAD	WATER TANK	<u> </u>		TIME			7	ACKING S	icica stad 11-17
DIGILLING EC	CME 75	334 SAMPLER	6/4 NEU		START			FINISH			
COMPLETION	н обрти	BEDROCK DEPTH	TOP OF CASING		9:30		9	:45			
I	17.0		3.0		AM			Am	se	CREEN 2	Pre, 5'. 01 5LOT
WELL CASIN	7.11	TYPE AND QUANTITY	- 1 1 1 m				S	AMI			PERSONNEL
	PUC SCH.	40, 5'SCREZ	7,10,3			-	_		_		L. Kan I Danas
		4						1 .	3	1	L. KEUN W. ROGERS
SCREEN INTE		TYPE AND QUANTITY				3	3	. 3	E 3	1 2 7	O. KEY BOSIE
ELEV.	12-17 ,5	PUC, OI SLOT		- 10	KTE	Sample	£ 8	Sample	Penetromi	N Value (Blows)	H-DALE HALFORD
ELEV.	ì	DESCRIPTIO	N		FERT	S	0 F	00 E	- E	Z 9	REMARKS
	17/150	10	111/2000 ZA		-						COM-CONTINIONS
	13.6-13.7	SHALL, COTRESE,	UNSDRIED, BROWN	`. E	J						
7		TINER SIZE AT	OUS SIZE, SURTEDIN	BOTTO				١.		ł	
		THE SIZE AT	80110-1		-/5-1						<u> </u>
-	IT & MB	7711 RO-101	1 11 71 0011	-	$\dashv$						X-TEMPORALY BENCHMARK
7	15.1-200	TILL, BR-GRA	y Sicily, Ence	F	コ		1	1			BENCHMARK
			SANDY		-16-	- 1	3	2			CHBLE TUTOVER
$\exists$	<b>.</b>			F	7	- [	V	٠,٥			
				E			İ				
$\cdot$ $\dashv$				17 -	17-		-	-			. 8 SANG BAKFEL
7				- E	コ						.0 3/1/4 0/1/20
$\dashv$					$\dashv$	- 1					
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1	BORING NO.	WELL HO.	GR	OUNOLEVEL E		¥				PAGE OF
1 -	COUNTY	WELL NO.  G103  SITE NO.		100.	53	} ·			I.R.	NULUS FILL MATERIAL
SIT	MARION	121050	0002	START			FINESH	AB		KING BENTONITE PELLETT
BORING LOC	SAMBOUAL	/ZINC		44/8	7	4	4/8	7		13-11
	WEST OF PL	HAT NEAR DITCH	TYPE		TIME			PA	CXING 5	JLICA SAND 11-17
ORCLING EQ		8 /4 WELL INSTALLE		START			THISH			
COMPLETION	17.01		3. o	11:30		0	30	sci	REEN 2	"PUC, 5'. OLSLOT
WELL CASING	TYPE	SCREEN, 5', 10'				S	AMP	LES		PERSONNEL
	7 . 9 . 901.10 . 7 . 3.		-							1. Keun W. ROSERS
SCREEN WYER		L 5' . 01 SLOT			No.	3	de recy Fi	Nomet	32	O-KEW BOSIE H-DALE HALFORA
ELEV.	12-17 PV	DESCRIPTION	70	нтчэ	Semp	Sample Type Sample Recover	Sample Recovery Penetron	Penetrome (Strength)	N Value (Blows)	REMARKS
			<u>-</u>	FEST	-			-		TEMARKS
			.	= =						*
크			E	- /]						
$\exists$			Е	3						1
크			E	-2-	1					AUGER DOWN
4			F	7					1	TO 10' WITH
7			E	2 ]				1		RELEASE FLITE
. =	- 100 LUREN	DOUT DARKBLIKE	4 - E	. J				ĺ		6/4
1	0-10,0 SURFACE	PEXT, DARKBLACK CRASS, TWISS, WET, OR	SHIK =	, =					1	
크	,	muck	=	4-	- 1					
7			F	_ ∃						
$\exists$			E	5-						
3			E	$\exists$						* TENDERAL MARK
				6						* TEMPORARY BENKE MARK CARLE TO TOWER
_ =	*		F	#						
7	, it	·	F	, ]					1	
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			=	<u>_</u> =						
				8-						
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$\dashv$			E	9-						
3			$\neg \nabla E$	3						
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=	- 0- /:-1/	44 24 20 21	F	· =				L	2	STOOM LEHGTY
7	0.2-12.3 SILTY	CLAY, TAN TO BROWN	F	,, ‡					2	3.8 .6 EXTENSION
$\exists$			F	(		2		Γ	8	
$\exists$			Е	_ =	1	उत्तर			10	
-]			E,	1	1			-	6	
1/2	3-14.3 KAND L	TGR-BROW, U.F. SHA	1 cen =	=		TUZO		+	_	
	J 4 / L	SATURATED		3-1	+'	4	- -	4	/	
7			F	7	1.			L	1.4	17.
٦			F	. ヿ	1.3	71 3	) [	1 .	- 11	100/2

### Illinois Environmental Protection Agency ABOVE PACKING BENTOXITE PETLETTS 4/4/87 10-11 PACKING SILICA SAND 11-17 6:30 11:30 SCREEN 2"PUC, S PM SAMPLES L. KEUN W. ROGERS O-KEH BOSIE H- DALE HALFORA 12-17 ELEV. DESCRIPTION REMARKS 24 TILL, GR.BR. PEOBLEY, SANDY STILL, DRY, CALC. 14.3-EOB 25 33 31 17.0 EOB. CABLE TO TOWER

### MONITOR WELL CONSTRUCTION

		Loc	ation: SANGUAL/ZINC
•			No.: 1210500002
Top of Protective Cover			II No.: 4/0/
Top of Casing3	3	(103.73) Prep	pared by: KEVIN W. ROLER
2		Depth (Elevation)	
	- 11	INFECT	
		1 100 15	, *
Ground Surface		(100.43	<del>-</del> )
Surface cap	N		
	14	()-	T
			Packed with
* TEMPORARY BENCH MARK 100.0	-		CEMENT GRUT
CONCRETE PAD FOR CABLE TU TOWER			7.5 BAGS .
NORTH EAST CORNER			TOP OF SEAL TO SURFACE 5% BENTONITE BY VOL
			3 % LENIONIE 134. VOL
		<b>5</b> .1	
		, 4	
	H	10.0 ( 90.43)+	Packed with
Bentonite seal	[Y]	11.0 (84 12)*	BENTONITE PELLETTS
	$\mathbb{N}$	N 11.0 1 84. 13 t	(.5 BUCKET
11.7		4.83 1 32.55*	
	コ目		Dankad with
Screen			Packed with
Total length 4.57			<u>SILICA SANO</u> 3.0 BAGS
Total length			J. 2. 1.7.
*			
		1 53.72	
Cap length	$\exists \Pi$	7	
Bottom of casing ———	$\dashv \sqcup$	16.7 ( 23.46)	
Bottom of boring		1 17.0 ( 83.43 1	* * * * * * * * * * * * * * * * * * * *
Duc	1		r / /montal /
Pipe: Type and quantity Prc 504 45, 5	<u> </u>	EN, 10 SECTUM,	S SCCIENT
TOTAL WELL LENGTH 20.02	. PK	DTECTIVE COVER 4"X	5'

#### MONITOR WELL CONSTRUCTION

E		Loca	tion: SAWDWAL/ZINC
		Site	No.: 12105 00002
Top of Protective Cover		(	No.:G102
Top of Casing	3.0		ared by: KEUN W. ROSER
		Depth (Elevation)	
			+
Ground Surface		( 98.54	-5
Surface cap	A	<b></b> ()	Γ-
F TEMFORTRY BENUT MARK 100.0 CONCREVE PAA FOR CABLE TV TOWER NORTA EAST CORNER			Packed with  CEMENT GROWT  5.5 BAGS  TOP OF SEAL TO SURFINE 3 & BONTONITE BY VOL.
Bentonite seal —-		<u>W35 88.19</u> )* 110 ( <u>87.54</u> *	Packed with  BENTONIE PELLETTS  1 BULLET
Screen Total length 4.64	12.0	(81.80)	Packed with  SIUCA SANA  GO BIBS
Cap length ————————————————————————————————————	#6	17.0 ( 81.54 *)	
Pipe: Type and quantity PYC SCH  TOTHL WELL LENGTH		LTCOLL GOR A"X	

#### MONITOR WELL CONSTRUCTION

		Location:	SAMLOVAL /=1.11
		Site No. :	1210500002
Top of Protective Cover		7() Well No.	: _ 6103
Top of Casing3.2	4	(100.53) Prepared	by: KellN W. Peter
		Depth (Elevation)	,
Ground Surface	7	( 47.53 <sup>*</sup>	
Surface cap		()	
TEMPORARI BENCH MARK 100.0  COM CRETE PAR FOR CABLE TV TOWER  ALORTH EAST CORNER		70	acked with  CEMENT GROUT  GO BAGS  OF OF SEAL TO SURFINE  CO BOLIONITY BY VCL.
Bentonite seal		, , , , ,	RENTONIE POLITS I BUCKET
<u>Screen</u>		-	cked with
Total length 4.55		1 80.78X	4.75 8465
Cap lengthBottom of casing		17.0 80.534	
Pipe: Type and quantity PVC ScH 40, 5's			SECTION
TOTAL WELL LENGTH 20.03, P	ROTE	ECTIVE CONFX 4 16.	

### MONITORING WELL INSTALLATION REPORT

. F. 14		INSTALLED: 6/14/90
BAN L		WELL NO. :
.7		BORING NO .:
DEPTH		PROJECT NO.: 8652.102
BELOW		PREPARED BY: 6. Webb
GROUND SURFACE	FTFT.	CHECKED BY :
FT.		TOP OF GROUND SURFACE EL. FT.
	TILSIIL TIBE	RISER PIPE EL. FT.
	GROUT MIX :	PROTECTIVE WELL COVER CASING :
		D K
		J N
		3 3
		RISER PIPE SCHEDULE
		ASTM DESIGNATION
2		1.0.1,8 0.0.2.0
Table #	BENTONITE SEAL:	COUPLINGS
T	1	PIPE IN 10, 03 FT. LENGTHS
	2 bentonite scal	V4.56 PIPE / 10.03 FT.
	using 1/2" bentonite	PIPE/4.00 FT.
	percets	SCREEN 5 FT. + .65 including point and
, 7/ _	. )	TOTAL 19.68 FT.
6,76 FT.		
8.76FT.		THICKNESS OF UPPER SEALFT.
		5.2
10,76 FT.		3.13
		15:
45	E:	LENGTH OF SCREEN 5 FT.
7		SLOT SIZE O, OLO IN.
7	· .	
	<u> </u>	
	[:]	LENGTH OF FILTER PACKFT.
	<u>[:]</u>	TYPE OF FILTER PACK
15.76FT.		
-3.12		1.52
18.0 FT.		BOTTOM OF BORING
-0-9		
REMAR	eks: El :	ath of point at base of screen
	565 to top	of the court at vace on several
	3 63 10 10 1	01
34		

## MONITORING WELL INSTALLATION REPORT

	INSTALLED: 6 15 90	15.
	WELL NO.: MWOZ	- 10
		- · · · · · ·
	BORING NO.:	-
DEPTH BELOW	PREPARED BY: K. Webb	•
GROUND	FT. FT. CHECKED BY:	_
SURFACE		_
FT.	7777-6:0:0:1	_FT.
	RISER PIPE EL. FT.	
2	GROUT MIX : PROTECTIVE WELL COVER CASING :	
		1887
	I N N N	5.45
	. RISER PIPE SCHEDULE	133
	ASTM DESIGNATION	
	1. D 0.D	
	BENTONITE SEAL:  COUPLINGS  PIPE INFT. LENGTHS	
	3' 1' bas 5 7 1	
	October 1	
	\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\	
	SCREENFT.	
9.30 FT.	FT.	
11,36 FT.	THICKNESS OF UPPER SEALFT.	
		•
13.35 FT.		
	LENGTH OF SCREENFT.	•
	SLOT SIZEIN.	ي ٩
		19.00 3.65
		13,3.5
	LENGTH OF FILTER PACKFT.	10.03
190 FT.	TYPE OF FILTER PACK	3.32
11.0 11.		
20,5 FT.	BOTTOM OF BORING	
REMAR	RKS:	
	· · · · · · · · · · · · · · · · · · ·	

# APPENDIX C ANALYTICAL RESULTS OF SUPPLEMENTAL FIELD INVESTIGATION

#### KEY TO TABLES

CRDL = Contract Required Detection Limits

B = Compound Detected in Laboratory Blank

J = Concentration is Estimated

NA = Compound Was Not Analyzed

U = Compound Was Not Detected

TABLE C-1
INORGANIC CONCENTRATIONS (MG/KG) IN TANK SAMPLES
SAMPLING LOCATIONS

ANALYTE	CRDL	TS01S	TS01D
AL LIMIT MILIM	200.0	39 U	70 11
ALUMINUM			38 U
ANTIMONY	60	1.9 U	1.8 U
ARSENIC	10	0.96 U	0.95 U
BARIUM	200	9.9 U	9.0 U
BERYLLIUM	5	0.97 U	0.95 U
CADMIUM	5	0.97 U	0.95 U
CALCIUM	5000	190 U	190 U
CHROMIUM	10	1.9 U	1.9 U
COBALT	50	9.9 U	9.5 U
COPPER	25	4.8 U	4.8 U
IRON	100	41	34
LEAD	5	28	28
MAGNESIUM	5000	190 U	190 U
MANGANESE	15	2.9 U	2.8 U
MERCURY	0.2	0.063 U	0.059 U
NICKEL	40	17	17
POTASSIUM	5000	99 U	91 U
SELENIUM	5	0.96 U	0.95 U
SILVER	10	1.9 U	1.9 U
SODIUM	5000	99 U	91 U
THALLIUM	10	1.9 U	1.8 U
VANADIUM	50	49	46
ZINC	20	20	19
CYANIDE	10	NA	NA
BTU		18,500	17,800

TABLE C-2
VOLATILE ORGANIC CONCENTRATIONS (UG/L) IN TANK SAMPLES
SAMPLING LOCATIONS

ANALYSIS	CRDL	TS01S	TS01D	
BENZENE	5.0	5000. U	5000. U	
TOLUENE	1.0	4400. J	6700.	
ETHYLBENZENE	5.0	20,000.	23,000.	
XYLENES (TOTAL)	10.0	96,000.	92,000.	

TABLE C-3
PCB CONCENTRATIONS (UG/KG) IN TANK SAMPLES
SAMPLING LOCATION

COMPOUND	CRDL	TS01S	TS01D
AROCLOR 1016	80.0	240,000 U	240,000 U
AROCLOR 1221	80.0	240,000 U	240,000 U
AROCLOR 1232	80.0	240,000 U	240,000 U
AROCLOR 1242	80.0	240,000 U	240,000 U
AROCLOR 1248	80.0	240,000 U	240,000 U
AROCLOR 1254	160.0	480,000 U	480,000 U
AROCLOR 1260	160.0	480,000 U	480,000 U

TABLE C-4
INORGANIC CONCENTRATIONS (MG/KG) IN SEDIMENT SAMPLES
SAMPLING LOCATIONS

ANALYTES	CRDL	ss01s	\$\$01D	\$\$02S	SS03S	\$\$04\$	SS04R .
ALUMINUM	200.0	18,000	12,000	9600	8560	13,600	200 U
ANTIMONY	60	3.4 U	3.1 U	4.0 U	3.1 U	16 U	10 U
ARSENIC	10	16	18	7.6	9.7	20	5.0 U
BARIUM	200	99	91	82	72	300	350
BERYLLIUM	5	1.6 U	1.4 U	1.8 U	1.5 U	2.2	5.0 U
CADMIUM	5	21	19	5.0	1.5 U	8.2	5.0 U
CALCIUM	5000	2500	2200	1700	1730	2960	12,600
CHROMIUM	10	18	3.8	22	3.1 U	13	10 U
COBALT	50	16 U	17	18 U	15 U	13 U	50 U
COPPER	25	820	850	440	330 J	1010 J	87
IRON	100	15,000	12,000	13,000	17,100 J	66,400 J	24,100 J
LEAD	5	2200	2000	490	190 J	2200 J	120
MAGNESIUM	5000	1700	1200	810	790	1720	1300
MANGANESE	15	260	1300	290	270	2770	170
MERCURY	0.2	0.17	0.14.	0.099 U	0.095 U	0.83	0.20 U
NICKEL	40	440	470	180	190	490	198
POTASSIUM	5000	890	530	390	490	560	29,200
SELENIUM	5	1.6 U	1.4 U	1.8 U	1.5 U	2.8	5.0 U
SILVER	10	3.2 U	2.9 U	3.6 U	31	46	10 U
SODIUM	5000	250	230	190	150 U	130 U	94,800
THALLIUM	10	3.4 U	3.1 U	4.0 U	3.1 U	2.7 U	27
VANADIUM	50	28	20	18 U	15 U	99	50 U
ZINC	20	18,000	15,000	150,000	1410 J	1080 J	16,200

TABLE C-5
PCB CONCENTRATIONS (UG/KG) IN SEDIMENT SAMPLES
SAMPLING LOCATIONS

COMPOUND	CRDL	SS01S	SS01D	\$\$02\$	SS03S	SS04S	
AROCLOR 1016	80.0	1400 U	1300 U	1700 U	1600 U	120 U	
AROCLOR 1221	80.0	1400 U	1300 U	1700 U	1600 U	120 U	
AROCLOR 1232	80.0	1400 U	1300 U	1700 U	1600 U	120 U	
AROCLOR 1242	80.0	1400 U	1300 U	1700 U	1600 U	120 U	
AROCLOR 1248	80.0	1400 U	1300 U	1700 U	1600 U	120 U	
AROCLOR 1254	160.0	2800 U	2500 U	3300 U	3200 U	230 U	
AROCLOR 1260	160.0	2800 U	2500 U	3300 U	3200 U	230 U	

TABLE C-6
INORGANIC CONCENTRATIONS (MG/KG) IN SURFACE SOIL SAMPLES
SAMPLING LOCATIONS

ANALYTE	 CRDL	\$\$05\$	SS05D	\$\$06\$	\$\$07\$	ss08s	\$\$09\$	\$\$10\$	SS10D	SS11S
ALUMINUM	200.0	11,300	11,000	7,520	6,000	6770	8990	7160	7,400	11,500
ANT IMONY	60	2.8 U	2.2 U	15 U	15	28	16	23	17	61
ARSENIC	10	13	6.7	7.9	21	57	55	12	21	38
BARIUM	200	200	130	84	150	190	200	290	270	130
BERYLLIUM	5	1.4	1.3 U	1.2 U	1.1 U	1.7	1.8	1.7	1.4	1.5 U
CADMIUM	5	1.4 U	1.3 U	1.2 U	1.1 U	3.7	21	67	35	1.5 U
CALCIUM	5000	10,900	3,500	96,800	14,300	1,830	3430	4570	2,440	1,640
CHROMIUM	10	2.8 U	5.1	2.4 U	24	9.5	16	98	18	60
COBALT	50	14 U	13 U	12 U	11 U	12 U	19	48	50	15 U
COPPER	25	190	73	100	350	4,290	4,250	5850	3,770	5,500
IRON	100	36,300	19,700	14,900	18,500	75,200	69,700	70,300	58,900	26,900
LEAD	5	510	130	250	4,000	41,000	16,000	11,000	6,200	21,000
MAGNESIUM	5000	3,310	1,290	5,460	300	590	550	1,180	480	830
MANGANESE	15	240	190	260	220	94	170	380	350	180
MERCURY	0.2	0.47	0.10 U	0.11	1.0	0.22	0.45	0.21	0.19	2.1
NICKEL	40	34	44	9.7 U	240	11	1,710	2,010	2,710	6D0
POTASSIUM	5000	1260	960	860	110 U	120 U	630	310	280	190
SELENIUM	5	1.7	1.3 U	1.2	1.1	2.8	2.8	1.6	1.7	3.6
SILVER	10	81	47	150	24	110	97	94	83	38
SODIUM	5000	220	130 U	310	- 110 U	200	470	380	280	190
THALLIUM	10	2.8 U	2.7 U	2.4 U	25	2.4 U	1.8 U	18	2.4 U	7.2
VANADIUM	50	14 U	13 U	12 U	11 U	43	110	97	77	15 U 1
ZINC	20	20,000	4200	2100	26,000	73,000	55,000	120,000	88,000	270,000

CRDL = CONTRACT REQUIRED

DETECTION LIMITS

S = SAMPLE

D = DUPLICATE

R = RINSATE

J = ESTIMATED VALUE

TABLE C-6 (CONTINUED)
INORGANIC CONCENTRATIONS (MG/KG) IN SURFACE SOIL SAMPLES
SAMPLING LOCATIONS

ANALYTE	CRDL	SS12S	SS12R	SS13S	SS13D	SS14S	SS14R	\$\$15\$ 	SS16S	SS16R	SS17S
ALUMINUM	200.0	11,800	200 U	12,900	10,300	8,600	200 U	11,100	10,400	200 U	8,810
ANTIMONY	60	19	60 U	13 U	2.4 U	2.6 U	60 U	2.2 U	150	60 U	25
ARSENIC	10	28	5.0 U	15	14	5.4	5.0 U	6.2	49	5.0 U	13
BARIUM	200	210	50 U	270	280	130	50 U	300	420	50 U	14 U
BERYLLIUM	5	1.3 U	5.0 U	1.1	1.6	1.0 U	5.0 U	1.2 U	1.8	5.0 U	1.4 U
CADMIUM	5	33	5 U	1.1 U	1,2 U	1.0 U	5 U	1.2 U	1.0 U	5 U	15
CALCIUM	5000	1,800	1,000 UJ	2030	990	980	1,000 UJ	630	5,210	1,000 UJ	1580
CHROMIUM	10	27	15	2.2	2.3 U	2.0 U	10	2.4 U	24	10 U	17
COBALT	50	13 U	50 U	11 U	12 U	10 U	50 U	12 U	10 U	50 U	14 U
COPPER	25	1350	53	490	520	100	35	6.0 U	2,880	25	3,000
I RON	100	35,300	66 J	36,500	42,100	15,900	50 UJ	22,300	57,600	64 J	37,400
LEAD	5	13,000	52	2,200	2,200	330	5 U	71	14,000	37	7,500
MAGNESIUM	5000	1,140	1,000 U	1,480	1,300	590	1000 U	800	490	1,000 U	750
MANGANESE	15	550	15 U	1,150	1,360	400	15 U	2,320	410	15 U	400
MERCURY	0.2	49	0.20 U	0.37	- 1.1	0.069 U	0.20 U	0.059 U	0.37	0.20 U	13
NICKEL	40	750	40 U	120	110	23	40 U	9.5 U	250	40 U	740
POTASSIUM	5000	260	500 U	730	530	470	500 U	640	480	500 U	140 U
SELENIUM	5	43	5.0 U	2.0	2.0	1.0 U	5.0 U	1.2 U	2.0	5.0 U	4.9
SILVER	10	54	10 U	64	67	23	10 U	43	91	10 U	42
SOD 1 UM	5000	140	500 U	730	640	100 U	500 U	200	550	500 U	140 U
THALLIUM	10	2.6 U	10 U	2.2 U	2.3 U	2.0 U	10 U	2.4 U	2.1 U	10 U	22
VANADIUM	50	13 U	50 U	26	27	10 U	50 U	24	76	50 U	14 U
ZINC	20	240,000	3,200	25,000	21,000	1,900	440	2200	24,000	100	210,000

CRDL = CONTRACT REQUIRED

DETECTION LIMITS

TABLE C-6 (CONTINUED)
INORGANIC CONCENTRATIONS (MG/KG) IN SURFACE SOIL SAMPLES
SAMPLING LOCATIONS

ANALYTE	CRDL	SS18S	SS19S	ss20s 	SS21S	\$\$22\$ —	SS23S	SS24S	ss25s	\$\$26\$	SS27S
ALUMINUM	200.0	6,130	6,530	7,770	10,700	7,310	6,540	5,710	8,910	9,630	6750
ANTIMONY	60	65	240	12 U	280	210 -	180	60	6.6	2.6 U	2.4 U
ARSENIC	10	1.1 U	28	12	31	18	23	35	9.5	8.7	5.3
BARIUM	200	11 U	160	75	290	120	68	74	89	150	120
BERYLLIUM	5	1.1 U	2.6	1.4	1.4	1.3	1.0	1.3 U	1.4 U	1.4 ປ	1.0 U
CADMIUM	5	48	27	1.0 U	27	10	3.7	1.3 U	1.4 U	1.4 U	1.0 U
CALCIUM	5000	4,180	23,500	1,670	29,200	2,090	4,500	750	960	480	1,270
CHROMIUM	10	13	73	4.4	8.3	14	16	2.6 U	2.7 U	2.7 U	2.0 U
COBALT	50	11 U	12 U	, 10 U	11 U	34	47	13 U	14 U	14 U	10 U
COPPER	25	1060	1,310	1,490	2,140	4,270	4,450	1,830	460	67	150
IRON	100	5380	126,000	32,100	56,600	44,600	54,100	43,200	16,300	21,000	18,400
LEAD	5	3200	6,300	1,300	7,600	4,300	14,000	28,000	830	170	15,000
MAGNESIUM	5000	2140	16,800	300	7,360	920	410	410	800	920	480
MANGANESE	15	340	3.5 U	4.2	320	120	290	13	390	1,790	910
MERCURY	0.2	1.4	5.7	0.45	0.66	0.53	0.46	7.7	0.11	0.098 U	0.081 U
NICKEL	40	490	450	780	1,410	2,500	3,460	600	240	25	66
POTASSIUM	5000	210	390	490	270	300	740	130 U	580	720	200
SELENIUM	5	1.1 U	9.6	1.4	1.5	1.3	2.5	5.5	1.4 U	1.4 U	1.0 U
SILVER	10	15	210	50	94	71	72	41	23	40	33
SODIUM	5000	160	180	260	130	200	620	130 U	140 U	140 U	100 U
THALLIUM	10	2.3 U	2.4 U	2.0 U	22	2.6 U	35	4.4	2.7 U	2.7 U	2.0 U
VANADIUM	50	11 U	100	10 U	78	66	95	13 U	14 U	17	10 U
ZINC	20	170,000	98,000	40,000	74,000	48,000	150,000	190,000	9,600	1,900	360,000

CRDL = CONTRACT REQUIRED

DETECTION LIMITS

TABLE C-7
PCB CONCENTRATIONS (UG/KG) IN SURFACE SOIL SAMPLES
SAMPLING LOCATION

PESTICIDES	CRDL	\$8088	ss10s	\$\$10D
AROCLOR 1016	80.0	1100 U	9600 U	9800 U
AROCLOR 1221	80.0	1100 U	9600 U	9800 U
AROCLOR 1232	80.0	1100 U	9600 U	9800 U
AROCLOR 1242	80.0	1100 U	9600 U	9800 U
AROCLOR 1248	80.0	1100 U	9600 U	9800 U
AROCLOR 1254	160.0	2100 U	19,000 U	20,000 U
AROCLOR 1260	160.0	2100 U	19,000 U	20,000 U

TABLE C-8
INORGANIC CONCENTRATIONS (MG/L) IN SURFACE WATER SAMPLES
SAMPLING LOCATIONS

ANALYTE	CRDL	sw01s	SW01D	SW02S	sw03s	SW04S	SW04R
ALUMINUM	200.0	780	1000	5,200	5,600	660	200 U
ANTIMONY	60	60 U	60 U	60 U	60 U	60 U	60 U
ARSENIC	10	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U
BARIUM	200	52	98	78	84	55	50 U
BERYLLIUM	5	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U
CADMIUM	5	360	370	5.0 U	5.0 U	5.0 U	5.0 U
CALCIUM	5000	100,000	110,000	17,000	5,300	18,000	1,000 U
CHROMIUM	10	10 U	10 U	10 U	10 U	10 U	10 U
COBALT	50	50 U	50 U	50 U	50 U	50 U	50 U
COPPER	25	90	80	85	90	79	25 U
IRON	100	1,400	1,400	3,200	3,300	3,200	50 U
LEAD	5	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U
MAGNESIUM	5000	16,000	16,000	2,400 .	2,500	4,400	1,000 U
MANGANESE	. 15	1,500	1,600	84 .	120	930	15 U
MERCURY	0.2	0.20 U	0.20 U	0.20 U	0.20 U	0.20 U	0.20 U
NICKEL	40	100	100	40 U	40 U	.40 U	40 U
POTASSIUM	5000	7,000	7,500	5,100	5,000	4,700	500 U
SELENIUM	5	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U
SILVER	10	120	120	17	10 U	13	10 U
SODIUM	5000	27,000	33,000	6,200	5,700	14,000	500 U
THALLIUM	10	53	47	10 U	10 U	10 U	10 U
VANADIUM	50	50 U	50 U	50 U	50 U	50 U	50 U
ZINC	20	4,200 J	4,100 J	500 J	1,000 J	1,000 J	110 J

TABLE C-9

VOLATILE ORGANIC CONCENTRATIONS (UG/L) IN
SURFACE WATER SAMPLES
SAMPLING LOCATIONS

VOLATILES	CRDL	sw01s	SW01D	SW02S	sw03s	sw04s	SW04R (2)
BENZENE	5.0	5 U	5 U	5 U	5 U	5 U	5 U
TOLUENE	1.0	5 U	5 U	5 U	25 B	5 B	27 B
ETHYLBENZENE	5.0	5 U	5 U	5 U	. 5 U	5 U	5 U
XYLENES (TOTAL)	10.0	10 U	10 U	10 U	10 U	10 U	10 U

TABLE C-10
INORGANIC CONCENTRATIONS IN RESIDENTIAL WELL
WATER SAMPLES
SAMPLING LOCATIONS

ANALYTES	CRDL	RW01S	200 U		
ALUMINUM	200.0	200 U			
ANTIMONY	60	60 U			
ARSENIC	10	5.0 U	5.0 U		
BARIUM	200	50 U	50 U		
BERYLLIUM	5	5.0 U	5.0 U		
CADMIUM	5	5 U	5 U		
CALCIUM	5000	66,900 J	68,800 J		
CHROMIUM	10	10 U	10 U		
COBALT	50	50 U	50 U		
COPPER	25	35	25 U		
IRON	100	2,700 J	2,660 J		
LEAD	5	5 U	5 U		
MAGNESIUM	5000	8,370	8,400		
MANGANESE	15	160	160		
MERCURY	0.2	0.20 U	0.2 U		
NICKEL	40	40 U	40 U		
POTASSIUM	5000	1,240	1,230		
SELENIUM	5	5.0 U	5.0 U		
SILVER	10	66	61		
SOD IUM	5000	12,300	12,400		
THALLIUM	10	10 U	110		
VANAD I UM	50	50 U	50 u		
ZINC	20	88	96		
CYANIDE	10	NA	NA		
BTU		NA	NA		

TABLE C-11
INORGANIC CONCENTRATIONS (UG/L) IN GROUNDWATER SAMPLES
SAMPLING LOCATIONS

NALYTES	CRDL	MW101S	MW101D	MW101R	MW102S	MW103s	MW01S TOTAL	MW01S DISSOLVED	MW02S TOTAL	MW02S D1SSOLVED
ALUMI NUM	200.0	200 U	200 U	200 U	200 U	200 U	14,000 J	500	13,000 J	200 11
YNOMITAN	60	60 U	60 U	60 U	60 U	60 U	7.0 U	7.0 U		200 U
RSENIC	10	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	17	7.0 U	7.0 U	7.0 U
BARIUM	200	50 U	50 U	50 U	50 U	50 U	160	50 U	16	5.0 U
BERYLLIUM	5	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	250	110
CADMIUM	5	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	45		5.0 U	5.0 U
CALCIUM	5000	24,000	24,000	1,000 U	240,000 J	290,000 J	1,100,000	32	6.0	7.0
CHROMIUM	10	10 U	10 U	10 U	10 U	10 U	150	1,100,000 75	130,000	110,000
COBALT	50	50 U	50 U	50 U	50 U	50 U	50 U	75 50 U	69	45
OPPER	25	25 U	25 U	25 U	25 U	25 U	64		50 U	50 U
RON	100	50 U	50 U	50 U	50 U	50 U	34,000 J	38	47	25 U
EAD	5	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	29	50 U	34,000 J	80
IAGNES I UM	5000	13,000	13,000	1,000 U	170,000	150,000	360,000 J	2.0 U	34	4.1
ANGANESE	15	15 U	15 U	15 U	15 U	380	1,500 J	380,000 J	46,000 J	38,000 J
ERCURY	0.2	0.20	0.33	0.20 U	0.20 U	0.2 U	0.20 U	710	1,400 J	270
ICKEL	40	40 U	40 U	40 U	40 U	40 U	41	0.20 U	0.20 U	0.20 U
OTASSIUM	5000	670	670	500 U	1,400	3,000	6,500	40 U	40 U	40 U
ELENIUM	5	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	3,000	5,900	2,200
ILVER	10	43	40	10 U	420	450	140	5.0 U 48	5.0 U	5.0 U
00 I UM	5000	240,000	241,000	500 U	280,000	95,000	420,000	430,000	11	10 U
HALLIUM	10	10 U	10 U	10 U	180	190	5.0 U	5.0 U	230,000	260,000
ANADIUM	50	50 U	50 U	50 U	50 U	50 U	50 U		5.0 U	5.0 U
INC	20	20 U	20 U	20 U	20 U	20 U	280	50 U	50 U	50 U
YANIDE	10	NA	NA	NA	NA .	NA NA		110	200	34
ITU		NA	NA	NA NA	NA NA	NA NA	NA NA	NA NA	NA NA	NA NA

TABLE C-12
VOLATILE ORGANIC CONCENTRATIONS (UG/L) IN
GROUNDWATER SAMPLES
SAMPLING LOCATIONS

VOLATILES	CRDL	MW01s	MW01T	MW02S	MW101s	MW101D	MW101R	
BENZENE	5.0	5 U	5 U	5 U	5 U	5 U	5 U	
TOLUENE	1.0	4 J	5 U	5 U	5 U	5 U	5 U	
ETHYLBENZENE	5.0	5 U	5 U	5 U	5 U	5 U	5 U	
XYLENES (TOTAL)	10.0	10 U	10 U	10 U	10 U	10 U	10 U	

TABLE C-13
PCB CONCENTRATIONS (UG/L) IN GROUNDWATER SAMPLES
SAMPLE LOCATIONS

PCBS	CRDL	MW01S	MW02S	MW101S	MW101D	MW101R	MW102S	MW103S
AROCLOR 1016	80.0	0.50 U	0.50 U	1.0 U	0.50 U	0.50 U	1.0 U	0.50 U
AROCLOR 1221	80.0	0.50 U	0.50 U	1.0 U	0.50 U	0.50 น	1.0 U	0.50 U
AROCLOR 1232	80.0	0.50 U	0.50 U	1.0 U	0.50 U	0.50 U	1.0 U	0.50 U
AROCLOR 1242	80.0	0.50 U	0.50 U	1.0 U	0.50 U	0.50 U	1.0 U	0.50 U
AROCLOR 1248	80.0	0.50 U	0.50 U	1.0 U	0.50 U	0.50 U	1.0 U	0.50 U
AROCLOR 1254	160.0	1.0 U	1.0 U	2.0 U	1.0 U	1.0 U	2.0 U	1.0 U
AROCLOR 1260	160.0	1.0 U	1.0 U	2.0 U	1.0 U	1.0 U	2.0 U	1.0 U

TABLE C-14
INORGANIC CONCENTRATIONS (MG/KG) IN WASTE PILE
AND ASH SAMPLES
SAMPLING LOCATIONS

			*						
ANALYSIS	CRDL	WPA01s	WPA01D	WPA02S	WPA03S	WPA04S	WPA05S	WPA06S	
ALUMINUM	200.0	74,000	79,000	37,000	27,000	10,000	15,000	1800	
ANTIMONY	60	46	42	8.1	9.6	20	6.6	6.8	
ARSENIC	10	14	17	29	26	29	28	9.3	
BARIUM	200	9.5 U	10 U	200	94	170	100	14	
BERYLLIUM	5	3.1	3.2	1.4	1.1 U	1.5	2.1	0.96 U	
CADMIUM	5	16	10	32	51	30	53	50	
CALCIUM	5000	690	910	11,000	3200	2400	3000	44	
CHROMIUM	10	330	300	330	110	40	55	1.9 U	
COBALT	50	44	44	62	12	13	11	9.6 U	
COPPER	25	210	160	71,000	1000	3800	590	210	
IRON	100	87,000	94,000	22,000	24,000	2300	62,000	3200	
LEAD	5	1100	1000	63,000	3200	8300	4300	10,000	
MAGNESIUM	5000	210	270	3000	2300	590	1500	120	
MANGANESE	15	56	58	1200	450	250	650	150	
MERCURY	0.2	0.065 U	0.059 U	0.25	0.70	0.063 U	0.30	0.059 U	
NICKEL	40	430	450	14,000	450	2000	430	49	
POTASSIUM	5000	780	830	140	2480	440	1360	170	*
SELENTUM	5	0.97 U	1.0 U	1.0 U	2.1	2.5	1.1	0.96 U	
SILVER	10	2.3	2.0	5.3	2.3 U	2.1 U	2.1 U	1.9 U	
SODIUM	5000	2610	2630	290	3530	570	1090	240	
THALLIUM	10	1.8 U	1.8 U	2.1 U	3.9	2.1 U	6.6	1.9 U	
VANADIUM	50	9.7 U	10 U	13	11 U	36	14	9.6 U	
ZINC	20	260,000	260,000	220,000	290,000	680,000	240,000	27,000	
CYANIDE	10	NA	NA	NA	NA	- NA	4A	NA	
BTU		NA							

TABLE C-15
EP TOXICITY CONCENTRATIONS (MG/KG) IN
WASTE PILE AND ASH SAMPLES
SAMPLING LOCATIONS

ANALYTE	CRDL	WPA01S	WPA01D	. WPA02S	WPA03S	WPA04S	WPA05S	WPA06S
			<del></del>	<del></del>	<del> </del>			·····
ARSENIC	10	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U
BARIUM	200	260	290	4000	1000	160	760	1200
CADMIUM	5	250	270	200	880	340	1500	210
CHROMIUM	10	10	8:6	8.7	6.9	12	7.2	6.2
LEAD	5	12	43	4000	8400	22,000	46,000	7100
MERCURY	0.2	0.20 U	0.20 U	0.20 U	0.20 U	0.20 U	0.20 U	0.20 U
SELENIUM	5	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U
SILVER	10	10 U	10 U	10 U	10 U	10 U	10 U	10 U

## TABLE C-16 VOLATILE ORGANIC CONCENTRATIONS (UG/L) IN GROUNDWATER SAMPLE SAMPLING LOCATIONS

VOLATILES	CRDL	MW01T	MW102S	MW103S	
CHLOROMETHANE	10.0	10. U	10. U	10. U	
BROMOMETHANE	10.0	10. U	10. U	10. U	
VINYL CHLORIDE	10.0	10. U	10. U	10. U	
CHLOROETHANE	10.0	10. U	10. U	10. U	
METHYLENE CHLORIDE	10.0	9. B	21. B	12. B	
ACETONE	10.0	10. U	10. U	10. U	
CARBON DISULFIDE	5.0	5. U	5. U	5. U	
1,1-DICHLOROETHENE	5.0	5. U	5. U	5. U	
1,1-DICHLOROETHANE	5.0	5. U	5. U	5. U	
1,2-DICHLOROETHENE	5.0	5.0	5. U	5. U	
CHLOROFORM	5.0	5. U	14. B	6. B	
1,2-DICHLOROETHANE	5.0	5. U	5. U	5. u	
2-BUTANONE	10.0	10. U	10. U	10. U	
1,1,1-TRICHLOROETHANE	5.0	5. U	5. U	5. U	
CARBON TETRACHLORIDE	5.0	5. U	5. U	5. U	
VINYL ACETATE	10.0	10, U	10. U	10. U	
BROMODICHLOROMETHANE	5.0	5. U	5. U	5. U	
1,2-DICHLOROPROPANE	5.0	5. U	5. U	5. U	
CIS-1,3-DICHLOROPROPENE	5.0	5. ບ	5. U	5. U	
TRICHLOROETHENE	5.0	5. U	5. U	5. U	
DIBROMOCHLOROMETHANE	5.0	5. U	5. U	5. U	
1,1,2-TRICHLOROETHANE	5.0	5. U	5. U	5. U	
BENZENE	5.0	5. U	5. U	5. U	
TRANS-1,3-DICHLOROPROPENE	5.0	5. U	5. U	5. U	
BROMOFORM	5.0	5. U	5. U	5. U	
4-METHYL-2-PENTANONE	5.0	10. U	10. U	10. U	
2-HEXANONE	10.0	10. U	10. ປ	10. U	
TETRACHLOROETHENE	5.0	5. U	5. U	5. U	
1,1,2,2-TETRACHLOROETHANE	5.0	5. U	5. U	5. U	
TOLUENE	1.0	5. U	5. U	2. J	
CHLOROBENZENE	5.0	5. U	5. U	5. U	
ETHYLBENZENE	5.0	5. U	5. U	5. U	
STYRENE	5.0	5. U	5. U	5. U	
TOTAL XYLENES	5.0	5. U	5. u	5. U	

TENTATIVELY IDENTIFIED

UNKNOWN

## APPENDIX D ESTIMATION OF VOLUME OF SOIL AND GROUNDWATER FOR REMEDIATION

## A. Estimation of Volume of Soil to be Remediated

The subsurface dimensions for the calculation were determined from the attached contour maps reproduced from the ISWS/ISGS report (Ref. 1). The surface dimensions were taken from the surveyed map presented in Appendix A, and also from the contour maps and average values used for calculations.

For the surveyed map, length = 9.1 inches From the scale of the map,  $\frac{1}{3}$  inch = 100 ft.

Length = 
$$\frac{4}{3}$$
 x 100 x 9.1 = 1,213 ft.

From contour map, length = 3 inches Scale,  $\frac{1}{3}$  inch = 300 ft.

Length = 
$$\frac{4}{3}$$
 x 300 x 3 = 1,200 ft.

Assume Average Length = 1,200 ft. (leaving allowance for some distance at the two ends).

From surveyed map, width = 
$$2.62 \times \frac{400}{3} = 350 \text{ ft.}$$

From contour map, width =  $1.25 \times 400 = 500 \text{ ft.}$ Average width = (350 + 500)/2 = 425 ft.

From Figure 14, lowest zinc concentration = 100 mg/kgAverage depth = (23 + 23 + 22 + 23 + 22 + 21 + 29 + 18)/9 = 22.5 ft.Volume to be remediated =  $(1,200)(425)(22.5)(1/27) = 425,000 \text{ yd.}^3$ 

From Figure 16, lowest cadmium concentration = 1 mg/kg Average depth = (17 + 18)/2 = (17.5)Volume to be remediated = (1,200)(425)(17.5)(1/27) = 330,555 yd.<sup>3</sup>

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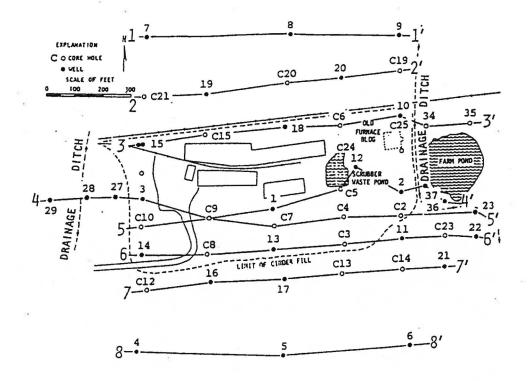
From Figure 17, lowest copper concentration = 100 mg/kgAverage depth = (7 + 3)/2 = 5 ft.Volume to be remediated =  $(1,200)(425)(5)(1/27) = 94,444 \text{ yd.}^3$ 

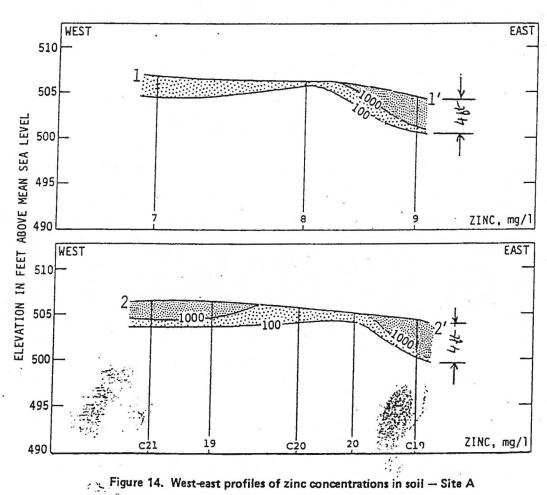
From Figure 18, lowest lead concentration = 100 mg/kgAverage depth = (8 + 7)/2 = 7.5 ft.Volume to be remediated =  $(1,200)(425)(7.5)(1/27) = 141,667 \text{ yd.}^3$ 

## B. Estimation of Volume of Groundwater to be Remediated

Assume average depth = 30 ft. Depth to water table = 5 ft. Net depth = (30-5) = 25 ft. Width = 425 ft., Length = 1,200 ft. Assume porosity = 15% Volume of groundwater below water table = (1,200)(425)(25)(0.15) ft.<sup>3</sup> (7.48 gal/ft<sup>3</sup>) = 14.3 x 10<sup>6</sup> gallons

Add 10% to account for volume of water above water table. Total volume to be remediated =  $(1.1)(14.3 \times 10^6) = 15.7 \times 10^6$  gallon





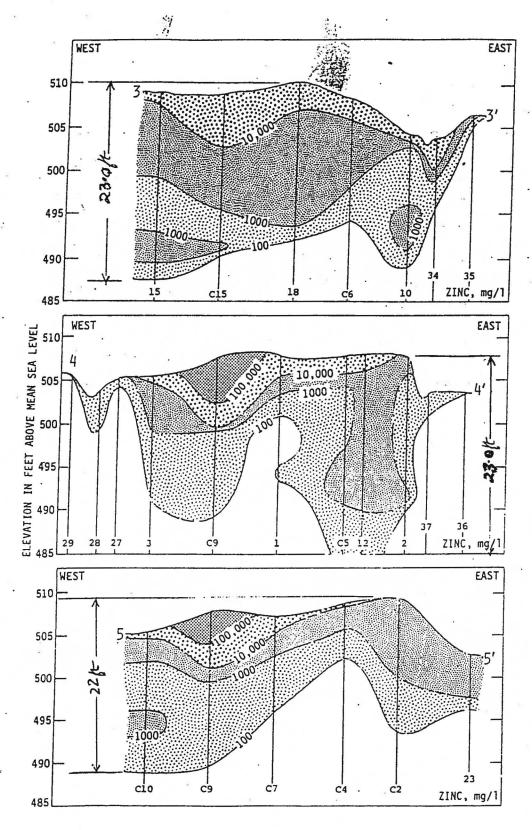


Figure 14. Continued

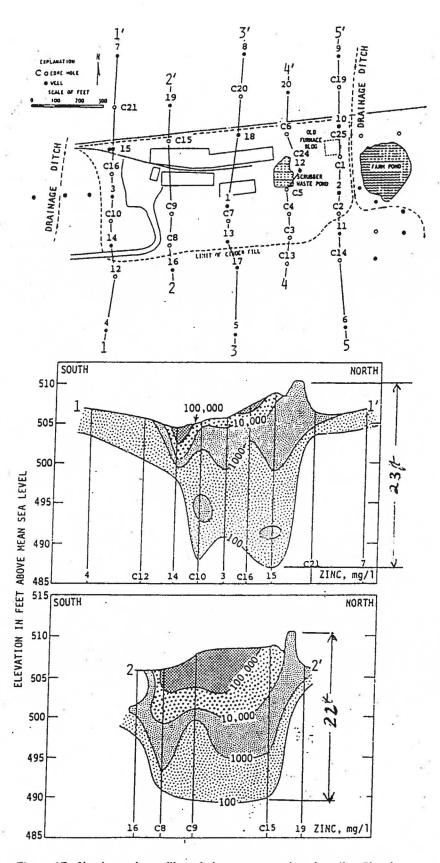


Figure 15. North-south profiles of zinc concentrations in soil - Site A

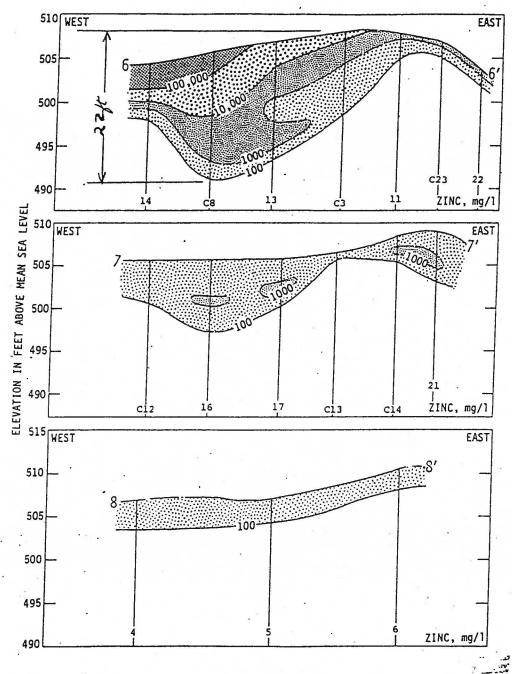


Figure 14. Concluded

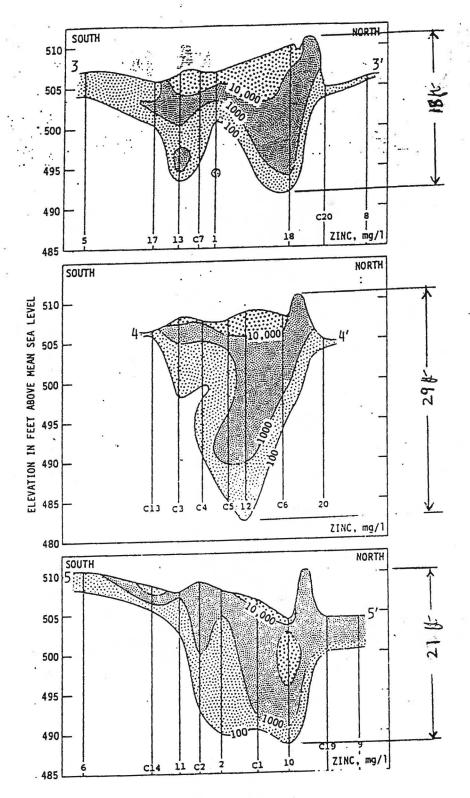
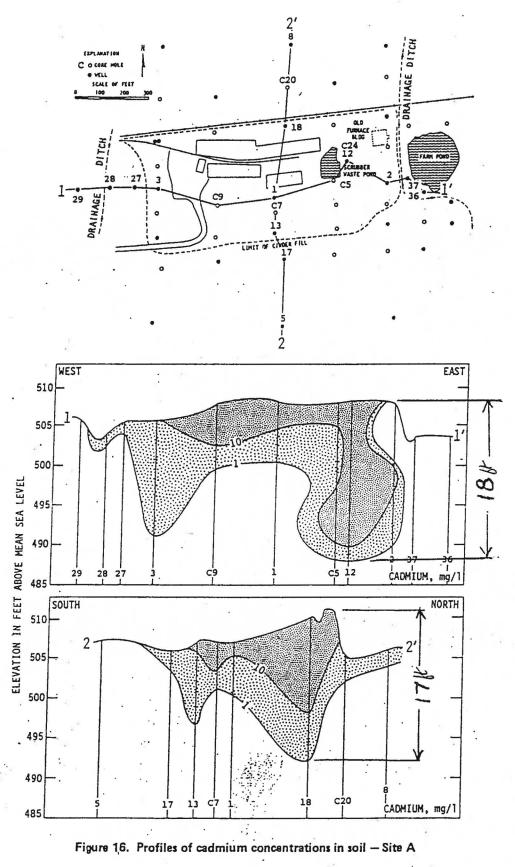


Figure 15. Concluded



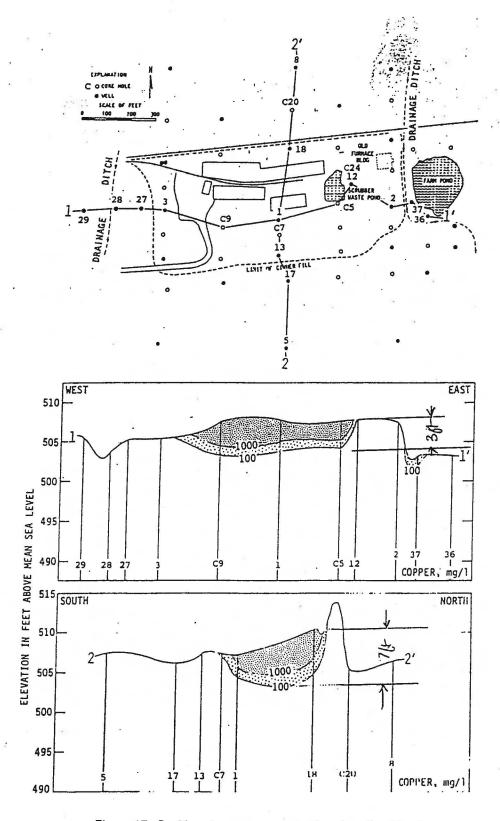


Figure 17. Profiles of copper concentrations in soil - Site A

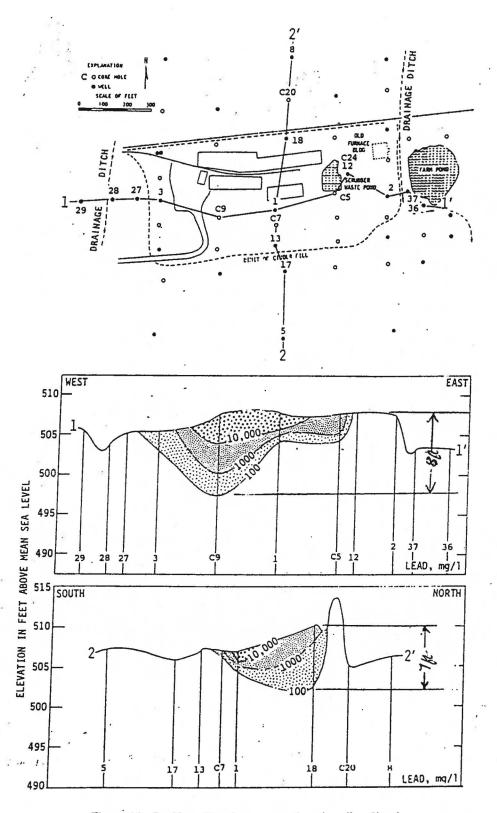


Figure 18. Profiles of lead concentrations in soil — Site A